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Abstract: Carbon stock and CO₂ emissions in agricultural systems are highly affected by the management of applied practices in arable farms, such as fertilizer use, irrigation, soil tillage, cover crop management, etc. This study evaluated the effects of various organic mulches, nitrogen fertilization and irrigation levels on soil CO₂ fluxes, soil carbon sequestration and processing tomato production in the Mediterranean environment. The field experiment was carried out with five main treatments, three cover crops of hairy vetch (HV), lacy phacelia (LF) and white mustard (WM) transplanted in autumn and cut in May to be used as mulches, plus barley straw mulch (BS) and conventional (C). After tomato transplanting, the main plots were split into two nitrogen fertilization treatments (0 and 100 kg N ha⁻¹) and the sub-plots were then split again into three irrigation levels (irrigation water 100%, 75%, 50% of evapotranspiration). In all treatments, a general effect was observed in the temporal fluctuations of soil CO₂ emissions throughout the observation period which were significantly influenced by soil temperature and water content. The temporal fluctuations of the soil CO₂ fluxes were attributed to climatic conditions and the peaks achieved optimal conditions of soil temperature and water content for soil respiration. A polynomial regression model highlighted the best conditions of soil temperature and water content in the various treatments in comparison (21.5 °C as average in BS, C and HV, and 18.3 °C in WM; soil water content of 28% as average in BS, C and HV, 55% in WM, and 87% in LP). A larger amount of TOC was observed in the mulching treatments than in the control after tomato harvesting, probably due to the residual biomass of the cover crops and a greater growth of the tomato. Although the soil carbon output as cumulated CO₂ emissions did not show statistically significant differences between the treatments, the soil carbon balance enabled us to estimate the highest net carbon contribution to the soil in HV determined by inputs and input/output ratio. However, except for the BS in 2013, the input/output ratios were >1 in all mulch treatments. In the Mediterranean environment, agronomical practices, such as the use of hairy vetch mulch on no-tilled soil, a slight reduction of irrigation water (-25%) and a rationalized use of N fertilizer potentially could shift the C balance in favor of soil C accumulation

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Dear Editor,

I am sending you the Manuscript titled “**Organic mulching, irrigation and fertilization affect soil CO₂ emission and C storage in the Mediterranean environment**” to be submitted for possible publication in Soil & Tillage Research. I hereby declare that the present Ms. is not either being submitted to or already published in another journal. All co-authors agree to submit the present Ms. version to Soil & Tillage Research.

The present Ms. Aims : (1) to evaluate the effects of various organic mulches, nitrogen fertilization and irrigation levels on soil CO₂ emissions; (2) to assess the impact of soil management with various organic mulches, nitrogen fertilization and irrigation levels on carbon sequestration; (3) to verify which type of organic mulching and which levels of nitrogen fertilization and irrigation are the most suitable for applying in processing tomato production in the Mediterranean environment. This is one of the few experimental studies carried out in field conditions where the impact of organic mulching, N fertilization and various irrigation levels are evaluated on soil CO₂ emissions, soil carbon content and processing tomato production in the Mediterranean environment. The main evidence provided in the study is that in the Mediterranean environment, environmental and agronomical benefits in agroecosystems could be obtained by increasing carbon inputs by choosing the most suitable agronomical practice such as the use of hairy vetch mulch on no-tilled soil, a slight reduction of irrigation water (-25%) and a rationalized use of N fertilizer which could potentially shift the C balance in favour of soil C accumulation.

Yours sincerely

Sara Marinari

Highlights

- fluctuations of soil CO₂ flux were related to soil temperature and water content
- strong effects on soil CO₂ emissions were observed on organic mulches from cover crops
- 25% water reduction appeared a good compromise for a positive soil carbon balance
- relations of soil temperature and water content with soil CO₂ flux were observed
- hairy vetch showed best effects on both tomato yield and carbon input/output ratio

1 **Organic mulching, irrigation and fertilization affect soil CO₂ emission**
2 **and C storage in the Mediterranean environment**

3

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17 **1. Introduction**

18 It is a well-known fact that the intensification of anthropic human activities increase greenhouse gas
19 (GHGs) in atmosphere with a consequent influence on climate (Canadell et al 2007). The
20 agriculture sector contributes significantly to the release of greenhouse gases into the atmosphere
21 (Paustian et al., 2004; Smith et al., 2008) which although it is only responsible for 6.9% of Italy's
22 total emission (Binatti et al., 2013), it plays an important role in mitigating strategies for reducing
23 GHG emissions. Consequently, there is a growing interest concerning the role of agriculture in
24 mitigating GHGs (Paustian et al., 1998; Mancinelli et al. 2008; Brahim et al. 2009). In agriculture
25 the carbon (C) stocks and the GHG emissions are highly affected by the management of the
26 agronomic practices applied in arable farms, such as fertilizer use, irrigation, soil tillage, cover crop
27 management, etc. (Mosier et al., 1998; Triberti et al., 2008; Kallenbach et al., 2010). In agricultural
28 cropping systems, most of the C is stored in the soil (Freibauer et al., 2004), therefore by means of
29 various soil practices, such as tillage, fertilization, mulching, crop residue incorporation and
30 irrigation (Lal 2004; Sainju et al., 2005; Gesch et al., 2007; Álvaro-Fuentes et al., 2008) it is
31 possible to increase soil organic matter (SOM) and then soil carbon content which is also one of the
32 principle parameters of soil quality (Andrews et al., 2002; Reicosky, 2008).

33 SOM plays an important role in the physical, chemical and biological properties of the soil (Manlay
34 et al., 2007) and represents the largest terrestrial pool in the C biogeochemical cycle (Houghton et
35 al., 2001). The increase of organic soil carbon stock is due to a positive budget between C inputs
36 (above and belowground biomass) and C outputs (heterotrophic soil organism and soil erosion)
37 (Kuzyakov, 2006). 120 Pg of atmospheric C-CO₂ are annually transformed into carbohydrates and
38 other organic molecules by plants with the natural process of photosynthesis and during the
39 decomposition process sequestered by SOC. The choice of the species and nutrient application can
40 positively affect the photosynthetic process thus increasing the net primary production and
41 consequently the carbon pools (Lal, 2006). The reduction of the number and intensity of soil tillage
42 directly (with less use of fossil fuel) and indirectly (with less alteration of soil characteristic and

43 fauna) affects soil CO₂ emissions and consequently soil C stock (Reicosky, 2008; West and
44 Marland, 2002). The results of several long-term experiments showed that the use of no-till (NT)
45 compared to conventional tillage (CT) can sequester nearly 60 g C m⁻² yr⁻¹ and soil tillage with
46 moldboard accelerates the SOM decomposition (West and Marland, 2002). Other studies in a corn–
47 soybean rotation showed that the use of less intensive tillage systems and better management of
48 crop residues are effective in reducing CO₂ emission and improving soil C (Al-Kaisi and Yin,
49 2005). The reduction of soil CO₂ emissions can be achieved by using agronomic practices that favor
50 soil C input and reduce the mineralization of soil organic matter, such as reduced soil tillage and
51 cropping system intensification which maximize water and nutrient use thus increasing production
52 (Paustian et al., 2000).

53 It is known that soil moisture, temperature and nutrient contents affect soil respiration and organic
54 matter decomposition (Kirschbaum, 2004). How soil moisture by irrigation and temperature affect
55 soil CO₂ emission in various herbaceous cropping systems is still an uncertain issue. A correct use
56 of irrigation water can promote both aboveground and root biomass production, favoring and
57 improving soil organic carbon concentration and decreasing C lost through erosion (Sauerbeck,
58 2001).

59 Nitrogen fertilization is one of the classical agronomic practices used in conventional cropping
60 systems with the aim of increasing yield and productivity. Concerning the soil CO₂ emissions and
61 soil C decrease, some studies show that nitrogen fertilization can negatively affect soil carbon stock
62 (Al-Kaisi et al., 2008; Mulvaney et al., 2009) while other studies show that fertilization has no
63 direct effect on soil carbon (Halvorson et al., 2002; Liang et al., 2012).

64 The introduction of cover crops in a crop rotation is an agronomic practice that are generally
65 adopted for addressing the improvement of cropping system sustainability and therefore soil quality
66 (Campiglia et al., 2009; Campiglia et al., 2011). A appropriate management of crop residues and the
67 use of winter cover crops can have a positive effect for limiting the leaching of residual Nitrogen
68 and increasing the nitrogen supply to the following main crop as well as improving both the

69 functions and properties of soil physical, chemical and biological, water movement, infiltration and
70 runoff (Clark et al., 1998; Liebig et al., 2002; Sainju et al., 2007; Mancinelli et al., 2013; Campiglia
71 et al., 2014). Specifically, the cover crop mulch residues left on soil surface can positively affect
72 various aspects of the cropping system such as the increase of water infiltration due to the decrease
73 of runoff, reduction of soil erosion, increase of moisture retention, weed control and reduction of
74 soil temperature in summer (Reicosky, 2008; Campiglia et al., 2012; Campiglia et al., 2014).

75 Although various studies have been carried out in order to evaluate the influence of soil tillage,
76 irrigation and N fertilization on some crops, little is known about their effects on soil CO₂ emissions
77 and vegetable crop production. Concerning vegetable cropping systems in Italy and in the
78 Mediterranean environment, to our knowledge no research has yet been carried out regarding the
79 impact of organic mulching, N fertilization and different irrigation levels on soil CO₂ emissions,
80 soil carbon content and processing tomato production yield.

81 It is assumed that, in the Mediterranean environment, a good combination of agronomical practices
82 such as the organic mulching of cover crops with irrigation and N fertilization can be profitably
83 used for growing vegetable crops in sustainable and/or organic agriculture. In order to find and
84 evaluate the best method for using organic mulching, irrigation and N fertilization on soil carbon
85 emissions and vegetable crop production yields in the Mediterranean environment, the processing
86 tomato was grown on organic mulch made from winter cover crops and barley straw.

87 The main objectives of this study were: (1) to evaluate the effects of various organic mulches,
88 nitrogen fertilization and irrigation levels on soil CO₂ emissions; (2) to assess the impact of soil
89 management with various organic mulches, nitrogen fertilization and irrigation levels on carbon
90 sequestration; (3) to verify which type of organic mulching and which levels of nitrogen
91 fertilization and irrigation are the most suitable for applying in processing tomato production in the
92 Mediterranean environment.

93

94 **2. Materials and Methods**

95 *2.1. Study site and soil characteristics*

96 The study was carried out in 2011/2012 and 2012/2013 in two adjacent experimental fields
97 established in September 2011 and in September 2012 at the experimental farm of the University of
98 Tuscia (Viterbo) located approximately 80 km North of Rome (45°25'N, 12°04'E).

99 The daily minimum and maximum temperatures and the rainfall during the study period were
100 collected by means of a meteorological station located 100 m from the study site. The climate of the
101 area is typical of the Mediterranean environment, generally characterized by cool, moist winters and
102 warm, dry summers. The long-term (last 30 years) mean annual amount of precipitation is 752 mm
103 of which 183 mm falls between June and September. The long-term mean annual average
104 temperature is 14 °C. The highest mean monthly temperatures occur between July and August when
105 the monthly maximum values are 31 °C and the monthly minimum values are 17 °C.

106 The soil of the experimental fields was classified as a *Typic Xerofluvent*, of volcanic origin. The
107 physicochemical characterization and sub-sequent classification analyses were carried out using the
108 official methods of analysis (MiPAF, 2000). The soil surface horizon (0–25 cm depth) was
109 classified as sand-loam USDA classification by the particle size distribution analysis, with 63%
110 sand, 22% silt, 15% clay. In the two adjacent experimental fields, the total organic C and N content
111 was on average 1.07 and 0.12, respectively (C/N ratio 9.2), pH(H₂O) of 7.1.

112 Chemical and physical analyses were carried out with twelve soil samples collected from the two
113 experimental fields (four soil samples in each block) before starting the trials (in September of both
114 years), in order to verify the field homogeneity. The soil properties did not differ within the three
115 blocks. In the year prior to the experimentation, barley (*Hordeum vulgare* L.) was grown in both
116 adjacent experimental fields.

117

118 *2.2. Experimental field and treatments*

119 The study focuses on the period of the growth of the tomato crop following a winter cover crop.

120 The area of the experimental field was 4400 m² (55 x 80) which made it possible to carry out all
121 farming operations with agricultural machinery. The experimental field was structured in five main
122 treatments arranged in a randomized block design with three replications. Three cover crops [hairy
123 vetch (*Vicia villosa* Roth, var. Villana) (hereinafter called HV), lacy phacelia (*Phacelia*
124 *tanacetifolia* Benth, var. Boratus) (hereinafter called LP) and white mustard (*Sinapis alba* L., var.
125 Emergo) (hereinafter called WM)] were sown in autumn and cut in spring to be used as mulches.
126 The other two main treatments consisted in barley straw mulch [hereinafter called BS] and
127 conventional (bare soil) [hereinafter called C].

128 In May the tomato plants were transplanted onto the cover crop mulches, barley straw mulch and in
129 the conventional plots. After tomato transplanting, the main plots were split into two nitrogen
130 fertilization treatments applied as: i) 0 kg N ha⁻¹ of nitrogen fertilization (hereinafter called N0); and
131 ii) 100 kg N ha⁻¹ of nitrogen fertilization (hereinafter called N100). The sub-plots were then split
132 again into three levels of irrigation: the amount of irrigation water ranging from 50% (hereinafter
133 called irr050), 75% (hereinafter called irr075) and 100% (hereinafter called irr100) of the total
134 evapotranspiration in the tomato crop.

135 The experimental design was a split-split-plot with three replications, where the main plots were
136 represented by the mulches [size 108 m² (18 m by 6 m)], the sub-plots were the nitrogen
137 fertilization treatments [size 54 m² (18 m by 3 m)], and the sub-sub-plots were the levels of
138 irrigation [size 18 m² (6 m by 3 m)].

139 In September of each year of the experiment, all plots were ploughed in at a depth of 30 cm,
140 fertilized with 100 kg of P₂O₅ ha⁻¹ as a triple superphosphate and then harrowed in at a depth of 10
141 cm in order to prepare the soil for the sowing of the cover crops. The cover crop species were sown
142 on September 26th, 2011 and September 30th, 2012 at seed rates of 60, 25 and 35 kg ha⁻¹ for HV,
143 LP, and WM, respectively. While, the soil of the other two treatments were kept bare and weed-free
144 throughout the cover crop growing season by chemical means (glyphosate applied when the weed
145 seedlings started to emerge). On May 8th, 2012 and May 14th, 2013, the cover crop aboveground

146 biomass was mowed using a hay-conditioner farm machine which cut the biomass to a width of 180
147 cm and arranged the residues in mulch strips about 80 cm wide and 100 cm apart. In the treatment
148 with barley straw mulch (no-tilled soil), the transplanting beds were prepared placing the straw in
149 strips at the rate of 400 g m⁻² of dry matter similar to those realized with cover crop aboveground
150 biomass. In the conventional treatment, the transplanting bed was prepared with plough and rotary
151 harrow at cover crop suppression. On May 18th, 2012 and May 27th, 2013, the tomato seedlings
152 (*Lycopersicon esculentum* Mill.) cv Ronco were hand-transplanted in paired rows into the mulch
153 layer at a distance of 40 cm between one another and a distance of 140 cm between the paired rows
154 at a density of 3 plants m⁻².

155 The drip irrigation tape, with 30-cm-spaced emitters, was installed over the mulch on each row of
156 tomato seedlings. The amount of water input was applied according to the irrigation treatment in
157 order to reintegrate 50%, 75% and 100% of maximum evapotranspiration estimated by a class A
158 evaporimeter and adjusted by crop coefficients (Allen et al. 1998). In the tomato plots where
159 nitrogen fertilization was foreseen, 50 kg N ha⁻¹ of ammonium nitrate was applied twice on June
160 15th and July 4th in the first year and on June 24th and July 9th in the second year. The weeds were
161 controlled whenever required by means of a rotary hoe between the paired rows and by hand inside
162 the paired rows.

163 The tomato was harvested on August 19th, 2012 and on August 20th, 2013.

164

165 2.3. Soil CO₂ emission, temperature and water content

166 In each of the 2 years, CO₂ emission from soil was measured weekly in each plot in the tomato-
167 growth period. In the event of rain the CO₂ emissions were measured at least 48 h after rainfall. The
168 measurements of soil respiration (SR) were carried out using a portable dynamic closed-chamber
169 infra-red gas analyzer system (Pumpanen, 2004). The CO₂ flux was measured between 8 am and 12
170 am in order to reduce diurnal variation in the CO₂ emitted without exception (Adviento-Borbe et al.,
171 2007). The non-steady-state through-flow chamber (SRC-1, PP Systems, Stotfold, UK; volume

172 1334 cm³, cover area 78.5 cm²) had only one opening to the soil and it was placed on a small
173 permanent PVC frame. The collars were placed in a central area of each plot between the paired
174 rows of the tomato crop.

175 The aboveground parts of the weeds were cut off before inserting the collars and clipped whenever
176 necessary. The increase of CO₂ within the chamber was monitored by means of a sensitive infra-red
177 gas analyzer instrument (EGM-4, PP Systems, Stotfold, UK). The closure time ranged between 30
178 and 180 seconds depending on the respiration rates of the soil. The SR was calculated automatically
179 by fitting a quadratic equation to the relationship between the increasing CO₂ concentration and
180 elapsed time. Each measurement took about 3 minutes.

181 The soil temperature close to the chamber was measured simultaneously to the measurement of CO₂
182 at a depth of 5 cm by using the “STP-1 Soil Temperature Probe” connected to an EGM-4
183 instrument.

184 The soil volume water content near the chamber was also measured simultaneously to the
185 measurement of CO₂ at a depth of 20 cm using the "TDR 300 Soil Moisture Meter" (Spectrum
186 Technologies, Inc., Plainfield, IL - USA).

187 In both years, the output of the soil CO₂ flux was estimated as the amount of CO₂ accumulated
188 (Reicosky, 1997; Curtin et al., 2000; Ding et al., 2007; Wilson and Al-Kaisi, 2008; Mancinelli et
189 al., 2010; Mancinelli et al., 2013) throughout the period of study. The calculations were made by
190 means of the linear interpolation of the two neighbouring measured fluxes and the numerical
191 integration over time (trapezoid rule) as reported in the following equation:

$$192 \quad \text{CO}_2 - \text{C} = \sum_i^n [(\mathbf{x}_i + \mathbf{x}_{i+1}) \times \mathbf{N} \div 2] + \dots + [(\mathbf{x}_{n-1} + \mathbf{x}_n) \times \mathbf{N} \div 2]$$

193 where: i = date of first measurement of CO₂ rate taken, n = date of last measurement of CO₂ rate
194 taken; x = CO₂ rate (kg ha⁻¹ day⁻¹), and N = number of days between the two consecutive CO₂ rate
195 measurements.

196

197 *2.4. Sampling and analysis of soil, cover crops and tomato*

198 In each year of the experiment the sampling of soil, cover crops and tomato was performed.

199 Before starting the experiment, five soil cores (0–20 cm depth) were taken from each plot and then
200 pooled together for the physicochemical characterization analysis. Just before the transplanting and
201 after the harvesting of the tomato crop, soil samples (0–20 cm depth) were collected from the
202 central inter-row of each plot, after removing the litter layer. The soil samples were air-dried, sieved
203 (<2 mm) and then kept at 4 °C in order to set up soil moisture and temperature at the most suitable
204 level for potential microbial activity (Zornoza et al., 2009).

205 The samples of cover crop aboveground biomass were collected in a 0.5 m² central area of each
206 plot just before mowing. The collected samples were dried at 70 °C until constant weight in order to
207 determine their dry weight and the C and N content.

208 The yield and aboveground biomass of the tomato crop were sampled in 2 m² of a central area of
209 each plot at crop physiological maturity. The samples of the yield and aboveground biomass of the
210 tomato were dried at 70 °C until constant weight in order to determine their dry weight and C and N
211 content.

212 Total organic carbon (TOC) and the total nitrogen (TN) contents of the soil, cover crops and tomato
213 were determined using an elementary analyzer (Thermo Soil NC—Flash EA1112).

214 Since a cover crop - tomato sequence can be considered a cropping system which is part of a wider
215 pluriannual crop rotation, belowground and aboveground biomass estimates of both cover crops and
216 tomato were used to calculate soil C input of the system. Empirical equations were used for
217 estimating the root residues derived from the C inputs of cover crops and tomato:

218 - Tomato roots (Mg dry wt. ha⁻¹) = 0.30 x aboveground biomass (Mg dry wt. ha⁻¹)
219 [adapted from Kong et al., 2005];

220 - Vetch roots (Mg dry wt. ha⁻¹) = 0.7 x aboveground biomass (Mg dry wt. ha⁻¹)
221 [adapted from Tian and Kang, 1998; Po et al., 2009];

222 - Phacelia roots (Mg dry wt. ha⁻¹) = 0.35 x aboveground biomass (Mg dry wt. ha⁻¹)
223 [adapted from Talgre et al., 2011];

224 - White m. roots (Mg dry wt. ha⁻¹) = 0.39 x aboveground biomass (Mg dry wt. ha⁻¹)

225 [adapted from Talgre et al., 2011].

226

227 *2.5. Data analysis and statistics*

228 The regression analyses of soil CO₂ evolution on soil temperature and volume water content was
229 performed for each treatment separately. The data used for the regression analyses of the soil CO₂
230 emission evolution on soil temperature and soil volume water content represented the mean value of
231 the two-year experiment. The effect of the soil temperature and the volume water content on the
232 CO₂ emission was plotted as a quadratic polynomial function. The results of the polynomial
233 regression analyses were useful for evaluating the influence of soil temperature and volume water
234 content on CO₂ emission.

235 For the data obtained in each experimental year, the differences in cover crop aboveground biomass
236 and relative C and N content, TOC and TON before tomato transplanting were statistically analyzed
237 by using the one-way ANOVA procedure. The data regarding the soil TOC and TON at tomato
238 harvesting, C input, C output, C input/output ratio of both experimental years were analyzed as a
239 split-split plot experimental design with the cover crop as main factor, the irrigation as the split
240 factor and the N fertilization as the split-split factor. Fisher's protected least significant differences
241 (LSD) at the 0.05 probability level (P<0.05) were used for comparing the main and interaction
242 effects.

243 In the manuscript the data of the soil CO₂ emission, soil temperature, and soil volume water content
244 are reported, which are presented as the means of the considered treatment ± standard error (SE).

245 Statistical analyses were performed using JMP statistical software package (SAS Institute, Cary,
246 NC).

247

248 **3. Results**

249 *3.1. Weather conditions*

250 From the historical data of the temperatures and precipitations observed in the same site and
251 recorded in the same agrometeorological station, the monthly average of the monthly minimum
252 temperatures were 10, 13, 16 and 17 °C in May, June, July and August, respectively and the
253 maximum temperatures were 23, 27, 31 and 31°C in May, June, July and August, respectively;
254 while, the total rainfall was 167 mm in the May-August period.

255 The minimum and maximum temperatures, and precipitation observed daily during the tomato
256 cultivation in the two years of experimentation are shown in figure 1. The trends of both
257 temperature and rainfall were considerably different in the 2012 and 2013 experimental years. The
258 2012 growing season of the tomato was generally hotter and drier compared to 2013. In fact, in the
259 May-August period of 2012 the maximum and minimum daily temperatures were lower than the
260 historical monthly values for 36 and 43 days, respectively; while in 2013 the maximum and
261 minimum daily temperatures were lower than the historical monthly values for 72 and 53 days,
262 respectively. In the two years experimental period, the amount of precipitation from May to August
263 was 110 and 172 mm in 2012 and 2013, respectively.

264

265 *3.2. Soil CO₂ emissions, temperature and volume water content*

266 The trends concerning the dynamics of soil CO₂ emission were quite similar throughout the study
267 period in both experimental years, with generally higher levels in July and lower levels at the end of
268 August. On average the value of soil CO₂ emission in July tended to be higher in 2013 than 2012
269 and in the first half of August of the second year there was a second peak of CO₂ in all treatments,
270 corresponding to the increase in air temperature during the same period. However, in August the
271 soil CO₂ emissions tended to be lower in all treatments, even if air temperatures remained rather
272 high while rainfall was absent. Nevertheless, the various mulching treatments showed different
273 trends of soil CO₂ emission in both experimental years (2012 and 2013) and the significant effects
274 of mulching treatments on soil CO₂ emission were evident throughout the study period in both years
275 (Fig. 2 and 3). In particular, the highest differences were observed among conventional, barley

276 straw mulch and hairy vetch treatments in both years of experimentation. These treatments were
277 influenced in different ways by the various climatic trends in the two-year experimentation. In fact,
278 the soil CO₂ emissions during the study period tended to be higher in HV, intermediate in BS and
279 lower in C in the first year, while, conversely, in the second year the soil CO₂ emissions tended to
280 be higher in C and lower in HV. The soil temperature was strongly related to the climatic trend
281 observed during the study period (Fig. 2 and 3). The highest significant differences between the
282 mulching treatments were observed in 2012 compared to 2013. The highest soil temperature values
283 (25.8 and 25.5 °C) were reached on August 20th, 2012 and July 10th, 2013 in the C and LP
284 treatments, respectively, while the lowest values (21.5 and 19.0 °C) were observed on July 27th,
285 2012 and June 28th, 2013 in the C treatment. The volume water content of the soil was generally
286 affected by the weather conditions (air temperatures and precipitation) during the measurement
287 period of both years (Fig. 2 and 3). Nevertheless, significant differences among treatments were
288 often observed, mainly in the first year of experimentation. From June onwards, the treatments with
289 mulched soil generally showed higher volume water content than C which was consequently the
290 most suitable management for determining lower levels of soil water retention, compared to the
291 mulched treatments.

292 The soil CO₂ emissions depend on soil temperature and volume water content and can be described
293 through the polynomial regression analysis for each of the mulching treatments (Fig. 4). With the
294 exception of the LP treatment, the polynomial regression model proved to be highly significant in
295 the relationship between soil temperature and soil CO₂ emissions in all other treatments (P= 0.0026,
296 P=0.0008, P=0.0163 and P=0.0052 for BS, C, WM and HV, respectively). The polynomial
297 regression curves show the highest level of soil CO₂ emission at 21.3, 21.6, 18.3 and 21.7 °C in BS,
298 C, WM and HV, respectively, and the increase or decrease in soil temperature determines a higher
299 reduction of soil CO₂ emissions in C compared to the other treatments (BS, WM and HV). There
300 was significant variation in the soil CO₂ emission response of soil volume water content, as
301 determined by the polynomial regression model for all mulching treatments (C: P<0.0001, BS:

302 P=0.0003, LP: P<0.0001, WM: P<0.0001, HV: P<0.0001). The polynomial regression curves
303 showed that all mulching treatments determine highest level of soil CO₂ values at different levels of
304 soil volume water content (25, 29, 87, 55 and 30% in C, BS, LP, WM, HV, respectively).

305 Concerning the irrigation treatment, soil CO₂ emission showed a similar dynamic to the mulching
306 treatment in the two-year study period. Throughout this period, significant differences were
307 observed between the three irrigation levels, with the highest level of soil CO₂ emission mainly in
308 the irr075 treatment, while the irr050 treatment often showed the lowest values (Fig. 5 and 6). In the
309 three irrigation levels, the temperature of soil was averagely related to the climatic conditions and
310 mainly to the daily maximum air temperature. Soil temperature generally increases according to the
311 decrease of quantity of irrigation water distributed. In fact, during the study period the lowest soil
312 temperature values were more often observed in the irr100, while the lowest values were observed
313 in the irr050. The lowest soil temperature values in the irr100 treatment were 21.8 and 18.4 °C
314 reached on July 27th, 2012 and June 28th, 2013, respectively. As expected, the soil volume water
315 content was averagely related to air temperature, while conversely the highest values were observed
316 for the soil temperature in the irr100 treatment, moreover these values were lower in the treatments
317 with reduced quantity of irrigation water distributed (Fig. 5 and 6).

318 In the relationship between soil CO₂ emissions and soil temperature, the multiple regression model
319 proved to be highly significant for the irrigation treatments irr050 and irr075 (P<0.0001), while the
320 100% irrigation (irr100) did not show a significant relationship (Fig. 7). The model showed the
321 highest level of soil CO₂ emissions at soil temperature of 22 °C for both levels of irrigation irr050
322 and irr075 but with CO₂ flux higher in irr075 than irr050 and a difference of 0.08 g m⁻² h⁻¹ of CO₂.

323 Soil CO₂ emissions plotted against soil volume water content were highly significant for all
324 irrigation treatments (P<0.0001). The model showed that the reduction of the quantity of irrigation
325 water (from 100% to 75% and 50%) causes a reduction of the soil water content required to
326 determine the maximum level of CO₂ emissions (43, 32 and 23% in irr100, irr075 and irr050,

327 respectively) and a decrease of CO₂ emitted (0.69, 0.65 and 0.58 g m⁻² h⁻¹ of CO₂ in the treatment
328 irr100, irr075 and irr050, respectively).

329 In relation to the nitrogen fertilization effect on soil CO₂ emissions, the results showed a general
330 trend similar to mulches and irrigation treatments throughout the measurement period. During the
331 study period, the un-fertilized treatment (N0) often showed higher values than the treatment
332 fertilized with 100 kg N ha⁻¹ (N100), in both experimental years. The soil temperature did not show
333 significant differences between N0 and N100, except at certain times of the second year. Soil
334 volume water content proved to be significantly influenced by nitrogen fertilization with lower
335 values in comparison to the un-fertilized treatment, mainly in July and August of both years (Fig. 8
336 and 9).

337 The polynomial regression model proved to be highly significant (P<0.0001) regarding soil CO₂
338 fluxes plotted against soil temperature and soil volume water content in both fertilized and
339 unfertilized treatments (Fig. 10). The curves show the highest CO₂ emissions of soil temperature for
340 fertilization treatments at 21 °C but with CO₂ flux lower in N100 compared to N0 (0.61 vs. 0.65 g
341 m⁻² h⁻¹ of CO₂ in N100 and N0, respectively). The soil CO₂ emissions related to volume water
342 content showed the highest values of CO₂ flux (0.62 vs. 0.64 g m⁻² h⁻¹ of CO₂ in N100 and N0,
343 respectively) at 30 and 33% of soil volume water content.

344

345 *3.3. Soil carbon and nitrogen content and crop productions*

346 At tomato harvesting in 2012 the conventional management showed the lowest values in soil TOC
347 and TON with 100% irrigation and in soil TOC with nitrogen fertilization, while in 2013 the
348 conventional treatment showed the lowest values in all fertilization and irrigation treatments (Tab.
349 1). In 2012 as well in 2013, the treatments with mulching showed on average higher soil TOC and
350 TON (1.44 and 0.14%, respectively) in comparison to the conventional treatment (1.33 and 0.13%
351 of TOC and TON, respectively), which was probably due to the biomass input. The soil TOC
352 tended to increase with irrigation at 100% (+0.01%) and nitrogen fertilization (+0.05%) in the

353 treatments with mulching, while it tended to decrease in the conventional treatment (-0.06 and -0.09
354 with irrigation at 100% and nitrogen fertilization, respectively). Higher rainfall and lower air
355 temperatures in 2013 compared to 2012 favored the carbon stock in the mulching treatment with
356 WM without nitrogen fertilization and with irrigation reduced to 50%. The higher air temperatures
357 which occurred in 2012 caused the highest levels of soil carbon accumulation in the HV mulching
358 without nitrogen fertilization and in the LP mulching with nitrogen fertilization at all irrigation
359 levels.

360 The value of HV aboveground biomass as mulching was higher in 2012 than 2013; conversely the
361 LP aboveground biomass was lower in 2012 than 2013. Nevertheless, the percentages of carbon in
362 the aboveground biomasses were similar in both years of experimentation. WM aboveground
363 biomass showed values significantly lower compared to LP and HV in 2012 as in 2013 (Tab. 2). In
364 the two experimental years, the N percentage in the HV biomass was, as expected, significantly
365 higher than WM and LP, determining a more favorable C/N rate for mineralization.

366 The higher amount of rainfall and lower air temperatures which occurred in 2013 generally tended
367 to reduce the yield of the tomato crop (-1.13 kg m^{-2} of FM), which was significantly affected by all
368 treatments applied (mulching, irrigation and nitrogen fertilization) (Tab. 2). Concerning the mulch
369 effect, the management with HV mulch produced a significantly higher tomato yield compared to
370 all other treatments in both experimental years (5.94 and 5.44 kg m^{-2} of FM). The WM mulch
371 produced the worst tomato yield (3.81 and 2.44 kg m^{-2} of FM). By reducing the quantity of
372 irrigation water from 100% to 75% and 50% there was a significant decrease in the tomato yield (-
373 15% and -49% with irr75 and irr50, respectively in 2012 and -12% and -48% with irr75 and irr50,
374 respectively in 2013). The nitrogen fertilization determined an increase in the yield of the tomato
375 crop in both experimental years (+42% and +57% in 2012 and 2013, respectively).

376

377 *3.4. Soil carbon balance*

378 The cover crop - tomato sequence can be considered to be a cropping system which is part of the
379 crop rotation of a farm. The higher carbon stock observed as average values in organic mulch
380 treatments (1.44% of soil TOC) compared to the conventional treatments (1.33% of soil TOC) (Tab.
381 1) is probably due to the result of variations of the C inputs in the two types of management.
382 Nevertheless, the amount of the soil CO₂ emissions was greater in the 2013 (4.05 Mg C ha⁻¹)
383 growing season when the soil emitted on average a significantly greater amount of CO₂ than in
384 2012 (3.40 Mg C ha⁻¹). In fact, significant differences were observed in carbon inputs between the
385 treatments of mulching, nitrogen fertilization and irrigation applied in both experimental years (Tab.
386 3). Conversely, the carbon outputs derived from CO₂ emissions showed no-significant differences
387 in the various types of management. The highest C inputs were observed in the HV mulch
388 management, while the conventional management showed the lowest carbon inputs in 2012 as in
389 2013. A net carbon depletion of the carbon balance was observed in the conventionally managed
390 system (input/output ratio <1) while there was an increase in the systems managed with organic
391 mulch from cover crops (1.4, 1.2, and 2.1 in LP, WM and HV respectively in 2012; 1.1, 1.1, 1.9 in
392 LP, WM and HV respectively in 2013). The BS mulch showed a disadvantage concerning the
393 input/output carbon rate in the second year. The nitrogen fertilization determined a significant
394 increase of carbon inputs in the system and a positive balance of input/output which was higher in
395 2012 than 2013. Significant differences of the applied irrigation treatments were only observed in
396 the second year concerning the input/output carbon rate for which the highest value was found in
397 the irr100.

398

399 **4. Discussion**

400 Many studies have been carried out and much is known regarding the beneficial effects on main
401 crops, soil quality, weeds and pest control produced by introducing cover crops into
402 agroecosystems. Although there are limited data and studies concerning the soil CO₂ emissions in

403 vegetable crops grown on organic mulches derived from previously-grown cover crops in the
404 Mediterranean environment.

405 The effect on soil CO₂ emissions was more pronounced in the tomato grown on organic mulches
406 made from the previous cover crop compared to the conventional management, which was more
407 significant in 2012 than 2013. Nevertheless, no significant statistical differences in cumulative CO₂
408 emissions were observed among the various agronomical techniques applied to the tomato
409 (mulching, irrigation, nitrogen fertilization), in accordance with other studies carried out in
410 agroecosystems (Wilson and Al-Kaisi, 2008; Mancinelli et al., 2010). A general effect was
411 observed in temporal variations of soil CO₂ emissions throughout the study period, in all mulching
412 treatments applied in 2012 as in 2013. Similarly, soil temperature and volume water content
413 variation during the study period were evident for all three agronomic techniques adopted.

414 The difference in soil CO₂ emissions observed in the mulching treatments could be possibly due to
415 the organic materials added in comparison to the conventional treatment, and also to the difference
416 in amount and quality of these organic materials among the various type of mulches, which had
417 different effects on the biological properties and organic carbon mineralization of the soil, in
418 accordance with other studies carried out on organic C mineralization (Collins et al., 2000; Iqbal et
419 al., 2009; Mancinelli et al., 2013). In fact, both cover crops and tomato produced a different
420 quantity of biomass as well as different qualitative characteristics of cover crop biomass between
421 one another and the two years of experimentation (Marinari et al., submitted paper). Several studies
422 demonstrated that soil carbon dynamics are highly influenced by plant biomass production.
423 Mancinelli et al (2010) found that the incorporation of cover crop biomass into the soil determined
424 an increase of soil CO₂ emissions. In another study, the switchgrass cultivation showed greater soil
425 CO₂ emissions than corn–soybean rotation, since twice the amount of crop biomass had been
426 incorporated into the soil (Al-Kaisi and Grote, 2007).

427 In this study the influence of soil temperature and volume water content on soil CO₂ emissions were
428 analyzed, since various studies have found that soil respiration may also depend on the temperature

429 and water content of the soil (Gaumont-Guay et al., 2006). The optimal values of soil temperature
430 and water content which cause the highest levels of activity of the biological processes for each
431 mulching treatment were estimated with a polynomial regression model, which revealed the optimal
432 conditions for soil respiration to be 21.5 °C as the average of the BS, C and HV, and 18.3 °C of the
433 WM. While the highest soil CO₂ emissions were estimated at a soil volume water content of 28% as
434 the average of the BS, C and HV, at 55% of the WM, and 87% of the LP. The significantly higher
435 levels of soil volume water content, generally observed in the mulching treatments rather than in
436 conventional treatments, enable us to deduce that the mulches applied increased soil water
437 availability which determined a more efficient use of the water, in agreement with Huang et al
438 (2003) and Hatfield et al. (2001). Therefore, the results of this study suggest that in the
439 Mediterranean environment agronomical management practices of soil tillage or no-tillage together
440 with organic mulch could cause variations in soil temperature and water content with consequent
441 beneficial effects on soil respiration.

442 The results of this study showed that soil CO₂ emission was positively affected by reducing the
443 quantity of irrigation water distributed to reintegrate the evapotranspiration to 75%, while the soil
444 CO₂ emission was negatively influenced when it was reduced to 50%. The increase of the quantity
445 of irrigation water (50%, 75% and 100% of the evapotranspiration) negatively influenced soil
446 temperature and positively influenced the soil volume water content as expected. However,
447 temporal variations in soil CO₂ emission, temperature and volume water content were observed at
448 all irrigation levels. In accordance with Sainju et al. (2006) and Jabro et al. (2008), by increasing the
449 quantity of irrigation water there was an increase in soil water content and consequently in soil
450 evaporation which probably reduced the soil temperature and therefore determined a higher soil
451 respiration reaching optimal levels. In fact, the polynomial regression model showed that the soil
452 CO₂ emission was higher in irr75 (0.67 g m⁻² h⁻¹) than in irr50 (0.59 g m⁻² h⁻¹) at the optimal soil
453 temperature of 22 °C.

454 Throughout the study period in the two experimental years the nitrogen fertilization (100 kg N ha⁻¹)
455 showed a reduction in soil CO₂ emissions except in July 2012; it is not surprising that soil
456 respiration decreased with the application of N fertilizer as observed in other studies (Lee and Jose,
457 2003; Ding et al., 2007; Al-Kaisi et al., 2008; Wilson and Al-Kaisi, 2008; Peng et al., 2011). In fact,
458 the polynomial model revealed a higher soil respiration in the fertilized treatment than in no-
459 fertilized at same optimum soil temperature (0.65 and 0.61 g m⁻² h⁻¹ in N100 and N0, respectively)
460 and similar values of soil volume water content (0.64 and 0.62 g m⁻² h⁻¹ in N100 and N0,
461 respectively).

462 After harvesting tomato, the greater soil TOC in the mulching treatments than in the control
463 treatments was probably due to the cover crop biomass and the greater growth of the total tomato
464 biomass. Although C and N contents in cover crop roots and in tomato roots were not actually
465 measured, some studies have shown that cover crop roots can provide up to 40% of the
466 aboveground biomass, influencing soil TOC and TON (Kuo et al., 1997a, 1997b; Sainju et al.,
467 2003) and therefore determining a major growth of the tomato plants (Sainju et al., 2001) and
468 increase of rhizodeposition of organic materials deriving from the tomato roots (Sainju et al., 2001)
469 causing a further increase of TOC and TON in the soil following tomato harvesting. Although the
470 total carbon was higher in the HV biomass than in LP and WM biomass in both years, soil TOC
471 after tomato harvesting was not always significantly higher compared to other treatments, probably
472 due to the its high N content and lower C/N rate, which facilitates the mineralization of biomass.

473 Teasdale et al. (2008) reported that hairy vetch biomass mineralizes rapidly after cutting
474 consequently releasing nitrogen which favors subsequent vegetable crops. In this study, although
475 different levels of production were observed in 2012 and 2013, the HV mulch determined the best
476 marketable fruit yields compared to the conventional treatment and the other mulching treatments
477 applied in both years of experimentation, in agreement with Campiglia et al. (2010, 2011).
478 Conversely, although environmental benefits in terms of soil CO₂ emission and soil carbon stock
479 were observed, the WM mulch provided the worst marketable fruit yield, as also reported also by

480 Hartz et al. (2005). A significant reduction in tomato yields was also caused by the lack of N
481 fertilization as well as the reduction of irrigation water. However, while the 50% reduction of
482 irrigation water determined a decrease of almost 50% of the tomato yield, the 25% reduction of
483 irrigation water showed a less than 15% decrease in yield.

484 In the concept of sustainable agriculture, agronomical techniques applied in the fields should be
485 seen from a holistic point of view as a system of management practices aimed at optimizing the soil
486 functions in agreement with Doran et al. (1994). Consequently, the cultivation of tomato on mulch
487 following a cover crop sequence as proposed in this study can be considered to be a cropping
488 system incorporated into a wider crop rotation. Therefore the soil carbon balance applied in this
489 study enabled us to estimate the carbon contribution derived from the proposed cropping systems
490 within a crop rotation. The soil carbon output intended as the cumulated CO₂ emissions over the
491 tomato growing season was not affected by any of the experimental treatments applied (mulching,
492 irrigation, N-fertilization). Conversely, the highest values were observed in the inputs as well as the
493 input/output ratios in the HV mulching while the lowest values were observed in C (<1). However,
494 the input/output ratios were always >1 in all mulch treatments, except the BS in 2013. The
495 reduction of irrigation determined a decrease in the input/output ratio, which was always >1 by
496 reducing the irrigation water to 75%. These results suggest that by applying organic mulches and/or
497 slightly reducing irrigation water (-25%) could allow the soil to become a temporary carbon sink
498 and therefore represent one of the possible solutions for reducing the soil CO₂ emissions in
499 sustainable agroecosystems, in agreement with Paustian et al. (2000) who proposed the agronomical
500 practices for increasing soil carbon inputs and reducing decomposition as the way for reducing the
501 net CO₂ emissions in the soil. The carbon input/output ratio >1 in the first year and <1 in the second
502 year generally caused by the lack of N-fertilization, made us assume that its rationalized distribution
503 in function of climatic conditions is important for increasing the sustainability of agricultural
504 systems.

505

506

507 **5. Conclusions**

508 This is one of the few experimental studies carried out in field conditions where the impact of
509 organic mulching, N fertilization and various irrigation levels are evaluated on soil CO₂ emissions,
510 soil carbon content and processing tomato production in the Mediterranean environment. According
511 to our results, the biomass quantity and qualitative characteristics of the cover crop appear to be the
512 main factors influencing the net soil CO₂ emissions and yield in the processing tomato grown on
513 organic mulches in no-tilled soil. There was a strong effect on soil CO₂ emissions in tomato grown
514 on organic mulching derived from the previous cover crops compared to the conventional, and the
515 highest influence was observed with the HV mulch, which determined higher positive values in
516 carbon balance in terms of input/output ratio. The 50% reduction of irrigation water caused both a
517 reduction in the soil CO₂ fluxes and in soil carbon inputs, therefore the 25% water reduction
518 appeared to determine a positive soil carbon balance. The nitrogen fertilization seems to slightly
519 reduce the CO₂ fluxes from soil but strongly increases the soil carbon inputs, therefore a rational
520 use of N fertilization in accordance with climatic conditions is required in order to increase the
521 sustainability of agricultural systems. However, the results highlighted that the fluctuations of soil
522 CO₂ emissions were related to the soil temperature and water content, and the estimated optimal
523 conditions for soil respiration differed according to the agronomical treatment applied.

524 In short it is possible to state that, although the cumulated CO₂ fluxes did not differ between the
525 various treatments applied, it is possible to achieve environmental benefits in terms of net soil CO₂
526 emissions and agronomical benefits in terms of processing tomato yield. In the Mediterranean
527 environment, environmental and agronomical benefits in agroecosystems could be obtained by
528 increasing carbon inputs by choosing the most suitable agronomical practice such as the use of
529 hairy vetch mulch on no-tilled soil, a slight reduction of irrigation water (-25%) and a rationalized
530 use of N fertilizer which could potentially shift the C balance in favor of soil C accumulation.

531

532

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538

539 **References**

- 540 Adviento-Borbe, M.A.A, Haddix, M.L., Binder, D.L., Walters, D.T., Dobermann, A., 2007. Soil
541 greenhouse gas fluxes and global warming potential in four high-yielding maize systems.
542 *Glob. Chang. Biol.* 13, 1972–1988.
- 543 Al-Kaisi, M.M., Yin, X., 2005. Tillage and Crop Residue Effects on Soil Carbon and Carbon
544 Dioxide Emission in Corn–Soybean Rotations. *Journal of Environmental Quality* 34 (2),
545 437-445.
- 546 Al-Kaisi, M.M., Grote, J.B., 2007. Cropping systems effects on improving soil carbon stocks of
547 exposed subsoil. *Soil Sci. Soc. Am. J.* 71, 1381–1388.
- 548 Al-Kaisi, M.M., Kruse, M.L., Sawyer, J.E., 2008. Effect of nitrogen fertilizer application on
549 growing season soil carbon dioxide emission in a corn-soybean rotation. *J. Environ. Qual.*
550 37 (2), 325–332.
- 551 Allen, R.G., Pereira, L.S., Raes, D, Smith, M., 1998. Crop evapotranspiration guidelines for
552 computing crop water requirements, irrigation and drainage. Paper No. 56. FAO Rome.
- 553 Álvaro-Fuentes, J., López, M. V., Cantero-Martinez, C., Arrúe, J.L., 2008. Tillage effects on soil
554 organic carbon fractions in Mediterranean dryland agroecosystems. *Soil Sci. Soc. Am. J.* 72,
555 541-547.

556 Andrews, S.S., Mitchell, J.P., Mancinelli, R., Karlen, D.L., Hartz, T.K., Horwath, W.R., Stuart,
557 Pettygrove, G., Scow, K.M. Munk, D.S., 2002. On-Farm Assessment of Soil Quality in
558 California's Central Valley. *Agron J.* 94, 12-22.

559 Binatti, R., Dramis, G., Angeloni, M., 2013. Sixth National Communication under the UN
560 Framework Convention on Climate Change: Italy. United Nations Framework Convention
561 on Climate Change,
562 http://unfccc.int/national_reports/annex_i_natcom/submitted_natcom/items/7742.php (last
563 access April, 15th, 2014)

564 Brahim, N., Gallali, T., Bernoux, M., 2009. Effect of Agronomic Practices on the Soil Carbon
565 Storage Potential in Northern Tunisia. *Asian J. Agric. Res.* 3, 55–66.

566 Campiglia, E., Paolini, R., Colla, G., Mancinelli, R., 2009. The effects of cover cropping on yield
567 and weed control of potato in a transitional system. *Field Crops Research* 112, 16-23.

568 Campiglia, E., Caporali, F., Radicetti, E., Mancinelli, R., 2010. Hairy vetch (*Vicia villosa* Roth.)
569 cover crop residue management for improving weed control and yield in no-tillage tomato
570 (*Lycopersicon esculentum* Mill.) production. *Europ. J. Agronomy* 33, 94-102.

571 Campiglia, E., Mancinelli, R., Radicetti, E., Marinari, S., 2011. Legume cover crops and mulches:
572 effects on nitrate leaching and nitrogen input in a pepper crop (*Capsicum annuum* L.).
573 *Nutrient Cycling in Agroecosystems* 89, 399-412.

574 Campiglia, E., Radicetti, E., Mancinelli, R., 2012. Weed control strategies and yield response in a
575 pepper crop (*Capsicum annuum* L.) mulched with hairy vetch (*Vicia villosa* Roth.) and oat
576 (*Avena sativa* L.) residues. *Crop Prot.* 33, 65–73.

577 Campiglia, E., Mancinelli, R., Di Felice, V., Radicetti, E., 2014. Long-term residual effects of the
578 management of cover crop biomass on soil nitrogen and yield of endive (*Cichorium endivia*
579 L.) and savoy cabbage (*Brassica oleracea* var. *sabauda*). *Soil & Tillage Research* 139, 1-7.

580 Canadell, J.G., Le Quere, C, Raupach, M.R., Field, C.B., Buitenhuis, E.T., Ciais, P., Conway, T.J.,
581 Gillett, N.P., Houghton, R.A., Marland, G., 2007. Contributions to accelerating atmospheric

582 CO₂ growth from economic activity, carbon intensity, and efficiency of natural sinks. Proc
583 Natl Acad Sci USA 104(47), 18866–18870.

584 Clark, M.S., Horwath, W.R., Shennan, C., Scow, K.M., 1998. Changes in Soil Chemical Properties
585 Resulting from Organic and Low-Input Farming Practices. *Agron. J.* 90, 662–671.

586 Collins, H.P., Elliott, E.T., Paustian, K., Bundy, L.G., Dick, W.A., Huggins, D.R., Smucker,
587 A.J.M., Paul, E.A., 2000. Soil carbon pools and fluxes in long-term corn belt
588 agroecosystems. *Soil Biol. Biochem.* 32, 157-168

589 Curtin, D., Wang, H., Selles, F., McConkey, B.G., Campbell, C.A., 2000. Tillage effects on carbon
590 fluxes in continuous wheat and fallow-wheat rotations. *Soil Sci. Soc. Am. J.* 64, 2080–2086.

591 Ding, W., Meng, L., Yin, Y., Cai, Z., Zheng, X., 2007. CO₂ emission in an intensively cultivated
592 loam as affected by long-term application of organic manure and nitrogen fertilizer, *Soil*
593 *Biol. Biochem.* 39, 669-679.

594 Ding, W., Meng, L., Yin, Y., Cai, Z., Zheng, X., 2007. CO₂ emission in an intensively cultivated
595 loam as affected by long-term application of organic manure and nitrogen fertilizer, *Soil*
596 *Biol. Biochem.* 39, 669-679.

597 Doran, J.W., Parkin, T.B., 1994. Defining and assessing soil quality. In: Doran, J.W., Coleman,
598 D.C., Bezdicek, D.F., Stewart, B.A. (eds) *Defining soil quality for a sustainable*
599 *environment*. SSSA, Spec Pub n°35, Madison, pp. 3-21.

600 Freibauer, A., Rounsevell, M.D.A., Smith, P., Verhagen, J., 2004. Carbon sequestration in the
601 agricultural soils of Europe. *Geoderma* 122, 1 –23.

602 Gaumont-Guay, D., Andrew Black, T., Griffis, T.J., Barr, A.G., Jassal, R.S., Nesic, Z., 2006.
603 Interpreting the dependence of soil respiration on soil temperature and water content in a
604 boreal aspen stand. *Agricultural and Forest Meteorology* 140, 220-235.

605 Gesch, R., Reicosky, D., Gilbert, R., Morris, D., 2007. Influence of tillage and plant residue
606 management on respiration of a Florida Everglades Histosol. *Soil Tillage Res.* 92, 156–166.

607 Halvorson, A.D., Wienhold, B.J., Black, A.L., 2002. Tillage, nitrogen, and cropping system effects
608 on soil carbon sequestration. *Soil Sci. Soc. Am. J.* 66 (3), 906–912.

609 Hartz, T.K., Johnstone, P.R., Miyao, E.M., Davis, R.M., 2005. Mustard cover crops are ineffective
610 in suppressing soilborne disease or improving processing tomato yield. *HortScience*, 40,
611 2016-2019.

612 Hatfield, J.L., Sauer, T.J., Prueger, J.H., 2001. Managing soils to achieve greater water use
613 efficiency: a review. *Agron. J.* 93, 271–278.

614 Houghton, J.T., Ding, Y., Griggs, D.J., Noguier, M., Van der Linden, P.J., Da, X., Maskell, K.,
615 Johnson, C.A., 2001. *Climate change 2001: the scientific basics*. Cambridge: Cambridge
616 University Press, New York, USA.

617 Huang, M., Dang, T., Gallichand, J., Goulet, M., 2003. Effect of increased fertilizer applications to
618 wheat crop on soil-water depletion in the Loess Plateau, China. *Agricultural Water
619 Management* 58, 267-278.

620 Iqbal, J., Hu, R., Lin, S., Ahamadou, B., Feng, M., 2009. Carbon dioxide emissions from Ultisol
621 under different land uses in mid–subtropical China. *Geoderma* 152, 63-73.

622 Jabro, J.D., Sainju, U., Stevens, W.B., Evans, R.G., 2008. Carbon dioxide flux as affected by tillage
623 and irrigation in soil converted from perennial forages to annual crops. *Journal of
624 Environmental Management* 88, 1478-1484.

625 Kallenbach, C.M., Rolston, D.E., Horwath, W.R., 2010. Cover cropping affects soil N₂O and CO₂
626 emissions differently depending on type of irrigation. *Agriculture, Ecosystems and
627 Environment* 137, 251–260.

628 Kirschbaum, M.U.F., 2004. Soil respiration under prolonged soil warming: are rate reductions
629 caused by acclimation or substrate loss? *Global Change Biology* 10, 1870-1877.

630 Kong, A.Y.Y., Six, J., Bryant, D.C., Ford Denison, R., and van Kessel, C., 2005. The Relationship
631 between Carbon Input, Aggregation, and Soil Organic Carbon Stabilization in Sustainable
632 Cropping Systems. *Soil Sci. Soc. Am. J.* 69, 1078-1085.

- 633 Kuo, S., Sainju, U.M., Jellum, E. J., 1997a. Winter cover crop effects on soil organic carbon and
634 carbohydrate. *Soil Sci. Soc. Am. J.* 61, 145-152.
- 635 Kuo, S., Sainju, U. M., Jellum, E. J., 1997b. Winter cover cropping influence on nitrogen in soil.
636 *Soil Sci. Soc. Am. J.* 61, 1392-1399.
- 637 Kuzyakov, Y., 2006. Sources of CO₂ efflux from soil and review of partitioning methods. *Soil Biol.*
638 *Biochem.* 38, 425–448.
- 639 Lal, R., 2004. Agricultural activities and the global carbon cycle. *Nutr. Cycl. Agroecosystems* 70,
640 103–116.
- 641 Lal, R., 2006. Carbon Management in Agricultural Soils. *Mitig. Adapt. Strateg. Glob. Chang.* 12,
642 303–322.
- 643 Lee, K.-H., Jose S., 2003. Soil respiration, fine root production, and microbial biomass in
644 cottonwood and loblolly pine plantations along a nitrogen fertilization gradient. *Forest*
645 *Ecology and Management* 185, 263-273.
- 646 Liang, C., Jesus, E.D.C., Duncan, D.S., Jackson, R.D., Tiedje, J.M., Balser, T.C., 2012. Soil
647 microbial communities under model biofuel cropping systems in southern Wisconsin, USA:
648 impact of crop species and soil properties. *Appl. Soil Ecol.* 54, 24–31.
- 649 Liebig, M., Varvel, G., Doran, J.W., Wienhold, B.J., 2002. Crop sequence and nitrogen fertilization
650 effects on soil properties in the western corn belt. *Soil Sci. Soc. Am. J.* 66, 596–601.
- 651 Mancinelli, R., Di Felice, V., Di Tizio, A., Lagomarsino, A., 2008. Carbon dioxide emission in
652 agricultural soils. In: Marinari S., Caporali F. (Ed) *Soil Carbon Sequestration Under Organic*
653 *Farming in Mediterranean Environment. Research Signpost*, 113-144.
- 654 Mancinelli, R., Campiglia, E., Di Tizio, A., Marinari, S., 2010. Soil carbon dioxide emission and
655 carbon content as affected by conventional and organic cropping systems in Mediterranean
656 environment. *Applied Soil Ecology* 46 (1), 64-72.

657 Mancinelli, R., Marinari, S., Di Felice, V., Savin, M.C., Campiglia, E., 2013. Soil property, CO₂
658 emission and Aridity Index as agroecological indicators to assess the mineralization of cover
659 crop green manure in a Mediterranean environment. *Ecological Indicators* 34, 31-40.

660 Marinari, S., Mancinelli, R., Brunetti, P., Campiglia, E. Soil quality, microbial functions and tomato
661 yield under cover crop mulching in Mediterranean environment. Submitted to *Soil Tillage
662 and Research*

663 Manlay, R.J., Feller, C., Swift, M.J., 2007. Historical evolution of soil organic matter concepts and
664 their relationships with the fertility and sustainability of cropping systems. *Agric. Ecosyst.
665 Environ.* 119, 217–233.

666 Mosier, A.R., Duxbury, J.M., Freney, J.R., Heinemeyer, O., Minami, K., 1998. Assessing and
667 mitigating N₂O emissions from agricultural soils. *Climatic Change* 40, 7-38.

668 Mulvaney, R.L., Khan, S.A., Ellsworth, T.R., 2009. Synthetic nitrogen fertilizers deplete soil
669 nitrogen: a global dilemma for sustainable cereal production. *J. Environ. Qual.* 38 (6), 2295–
670 2314.

671 Paustian, K., Cole, C., Sauerbeck, D., Sampson, N., 1998. CO₂ mitigation by agriculture: an
672 overview. *Clim. Change* 40, 135–162.

673 Paustian, K., Six, J., Elliott, E.T., Hunt, H.W., 2000. Management options for reducing CO₂
674 emissions from agricultural soils. *Biogeochemistry* 48, 147-163.

675 Paustian, K., Babcock, B., Kling, C., Hatfield, J., Lal, R., Mccarl, B., Mclaughlin, S., Post, W.M.,
676 Mosier, A., Rice, C., Robertson, G.P., Rosenberg, N.J., Rosenzweig, C., 2004. *Agricultural
677 Mitigation of Greenhouse Gases: Science and Policy Options, Framework.*

678 Peng, Q., Dong, Y., Qi, Y., Xiao, S., He, Y., Ma, T., 2011. Effects of nitrogen fertilization on soil
679 respiration in temperate grassland in Inner Mongolia, China. *Environmental Earth Sciences*
680 62, 1163-1171.

681 Po, E.A., Snapp, S.S., Kravchenko, A., 2009. Rotational and Cover Crop Determinants of Soil
682 Structural Stability and Carbon in a Potato System. *Agron. J.* 101, 175-183.

683 Pumpanen, J., 2004. Comparison of different chamber techniques for measuring soil CO₂ efflux.
684 Agric. For. Meteorol. 123, 159-176.

685 Reicosky, D.C., 1997. Tillage-induced CO₂ emission from soil. Nutr. Cycl. Agroecosys. 49, 273-
686 285.

687 Reicosky, D., 2008. Carbon sequestration and environmental benefits from no-till systems, In:
688 Goddard, T.e.a. (Ed.), No-till Farming Systems. World Association of Soil and Water
689 Conservation, Bangkok, pp. 43-58.

690 Sainju, U. M., Singh, B.P., Whitehead, W.F., 2001. Cover crops and nitrogen fertilization effects on
691 soil carbon and nitrogen and tomato yield. Canadian Journal of Soil Science 80(3), 523-532.

692 Sainju, U. M., Singh, B.P., Whitehead, W.F., 2001. Comparison of the effects of cover crops and
693 nitrogen fertilization on tomato yield, root growth, and soil properties. Scientia Horticulturae
694 91, 201-214.

695 Sainju, U. M., Whitehead, W.F., Singh, B.P., 2003. Cover crops and nitrogen fertilization effects on
696 soil aggregation and carbon and nitrogen pools. Canadian Journal of Soil Science 83(2),
697 155-165.

698 Sainju, U.M., Singh, B.P., Whitehead, W.F., 2005. Tillage, cover crops, and nitrogen fertilization
699 effects on cotton and sorghum root biomass, carbon, and nitrogen. Agron. J. 97, 1279.

700 Sainju, U.M., Jabro, J.D., Stevens, W.B., 2006. Soil carbon dioxide emission as influenced by
701 irrigation, tillage, cropping system, and nitrogen fertilization. In: Aneja, V.P., Schlesinger,
702 W.H., Knighton, R., Jennings, G., Niyogi, D., Gilliam, W., Duke, C.S., (Eds.), 2006.
703 Proceedings of Workshop on Agricultural Air Quality: State of the Science, Maryland,
704 USA, pp. 1086-1098

705 Sainju, U.M., Singh, B.P., Whitehead, W.F., Wang, S., 2007, Accumulation and crop uptake of soil
706 mineral nitrogen as influenced by tillage, cover crops, and nitrogen fertilization. Agron J 99,
707 682-691.

708 Sauerbeck, D.R., 2001. CO₂ emissions and C sequestration by agriculture: perspectives and
709 limitations. *Nutr. Cycl. Agroecosystems* 253–266.

710 Smith, P., Martino, D., Cai, Z., Gwary, D., Janzen, H., Kumar, P., McCarl, B., Ogle, S., O'Mara, F.,
711 Rice, C., Scholes, B., Sirotenko, O., Howden, M., McAllister, T., Pan, G., Romanenkov, V.,
712 Schneider, U., Towprayoon, S., Wattenbach, M., Smith, J., 2008. Greenhouse gas mitigation
713 in agriculture. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* 363, 789–813.

714 Talgre, L., Lauringson, E., Makke, A., Lauk, R., 2011. Biomass production and nutrient binding of
715 catch crops. *Agriculture*, 98 (3), 251-258.

716 Teasdale, J.R., Abdul-Baki, A.A., Bong Park, Y., 2008. Sweet corn production and efficiency of
717 nitrogen use in high cover crop residue. *Agron. Sustain. Dev.* 28, 559-565.

718 Tian, G., Kang, B.T., 1998. Effects of soil fertility and fertilizer application on biomass and
719 chemical compositions of leguminous cover crops. *Nutr. Cycl. Agroecosyst.* 51, 1385-1314

720 Triberti, L., Nastri, A., Giordani, G., Comellini, F., Baldoni, G., Toderi, G., 2008. Can mineral and
721 organic fertilization help sequester carbon dioxide in cropland? *Eur. J. Agron.* 29, 13–20.

722 West, T.O., Marland, G., 2002. A synthesis of carbon sequestration, carbon emissions, and net
723 carbon flux in agriculture: comparing tillage practices in the United States. *Agric. Ecosyst.*
724 *Environ.* 91, 217–232.

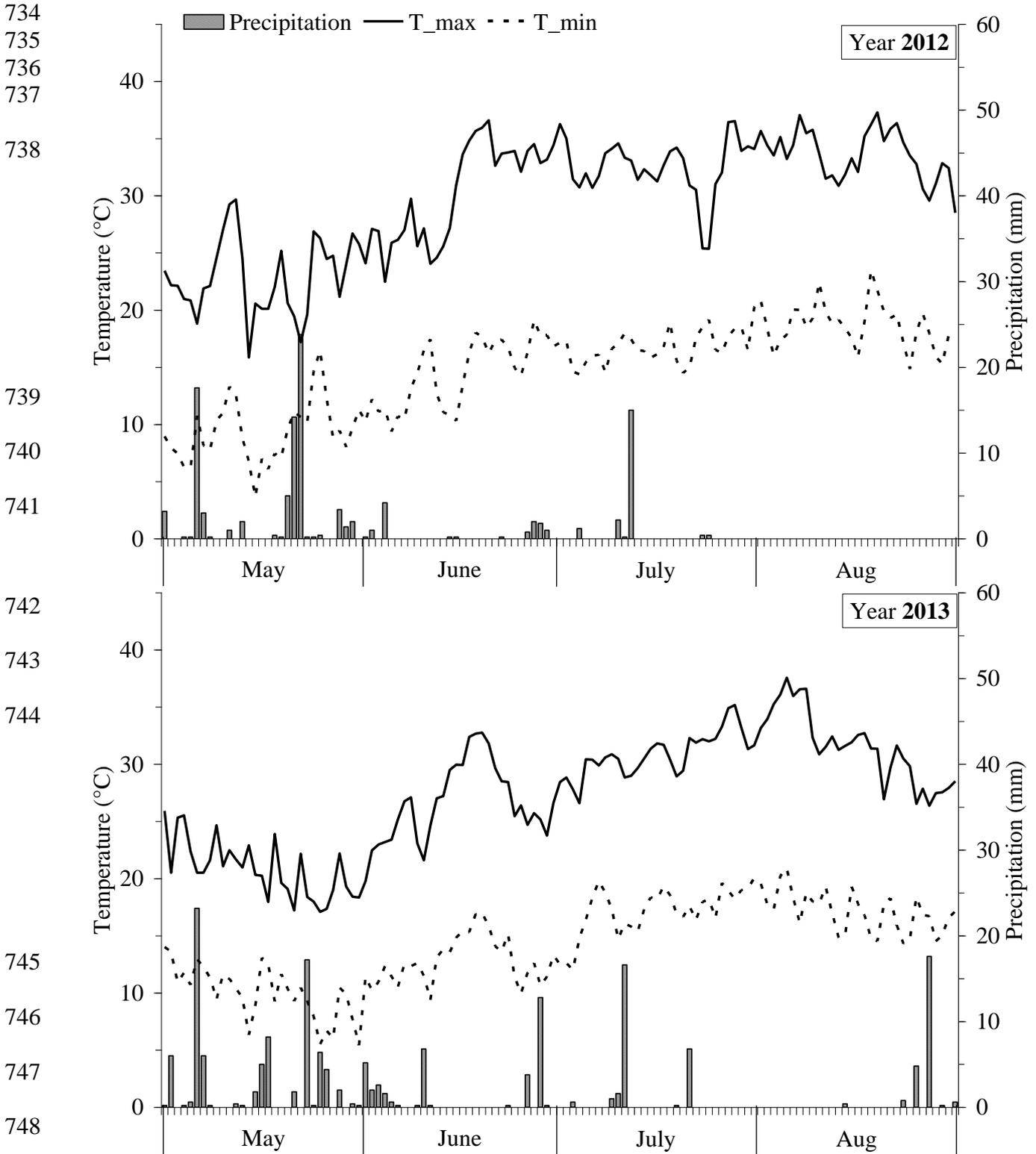
725 Wilson, H.M., Al-Kaisi, M.M., 2008. Crop rotation and nitrogen fertilization effect on soil CO₂
726 emissions in central Iowa. *Appl. Soil Ecol.* 39, 264-270.

727 Zornoza, R., Mataix-Solera, J., Guerrero, C., Arcenegui, V., Mataix-Beneyto, J., 2009. Storage
728 effects on biochemical properties of air-dried soil samples from Southeastern Spain. *Arid*
729 *Land Research and Management* 23 (3), 213–222.

730

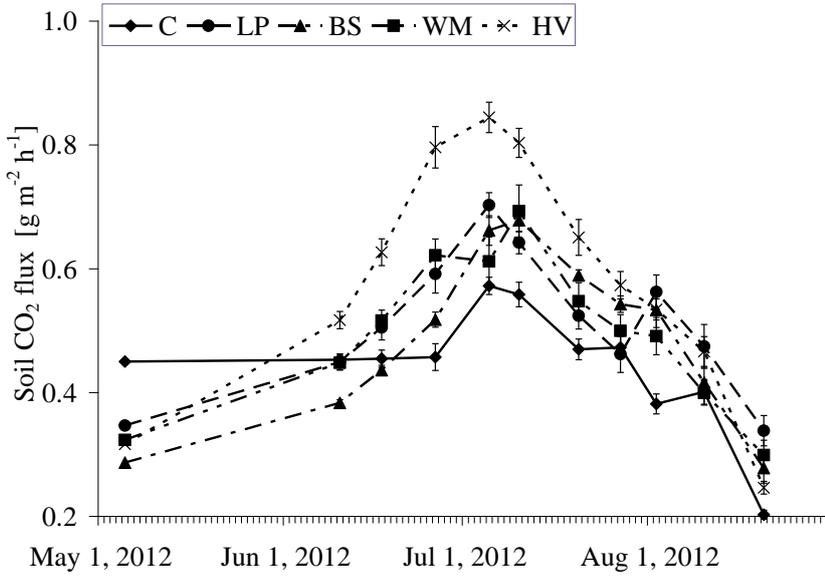
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732 Figure 1. Daily minimum [---] and maximum [—] temperatures (°C), and rainfall [□] (mm) at the
733 experimental site, throughout the periods of study in 2012 and 2013.

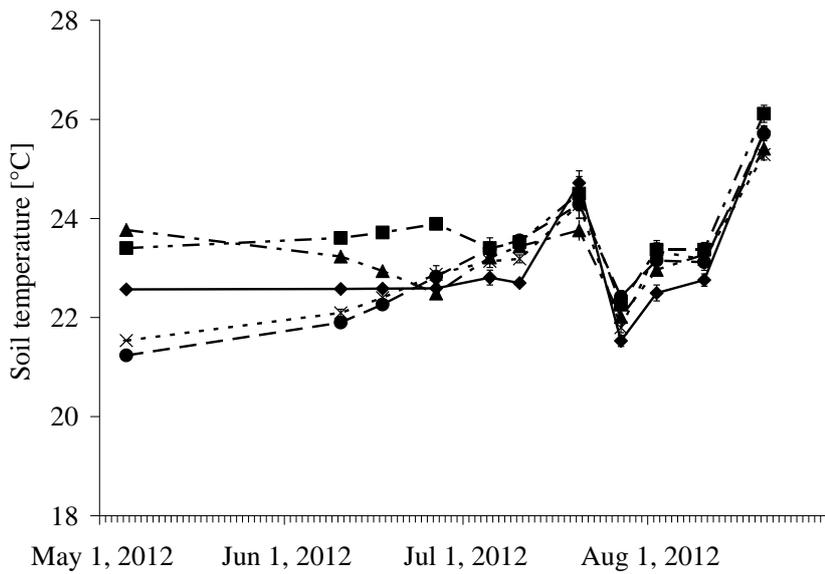


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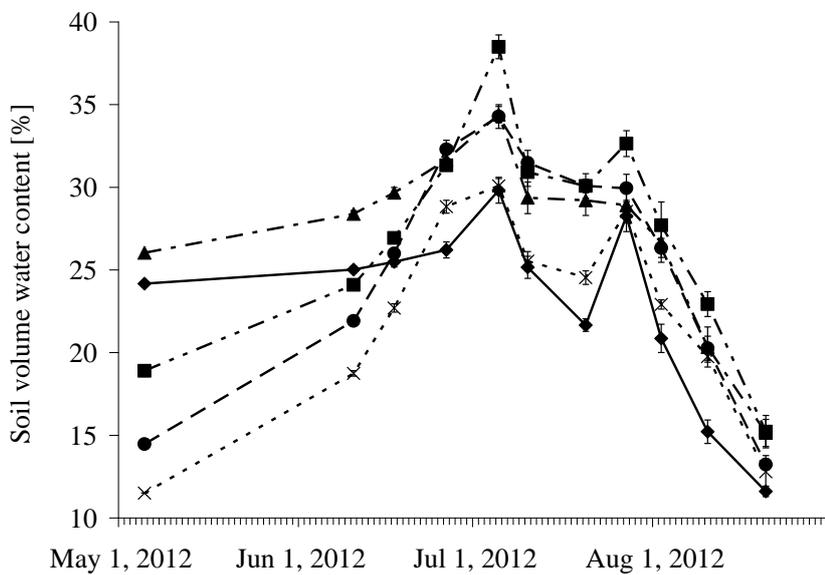
752 Figure 2 - Soil CO₂ flux, soil temperature and soil volume water content in the various mulching
 753 treatments, during the 2012 tomato cycle. Bars are standard errors (n = 18). BS: barley straw; C:
 754 conventional; HV: hairy vetch; LP: lacy phacelia; WM: white mustard.



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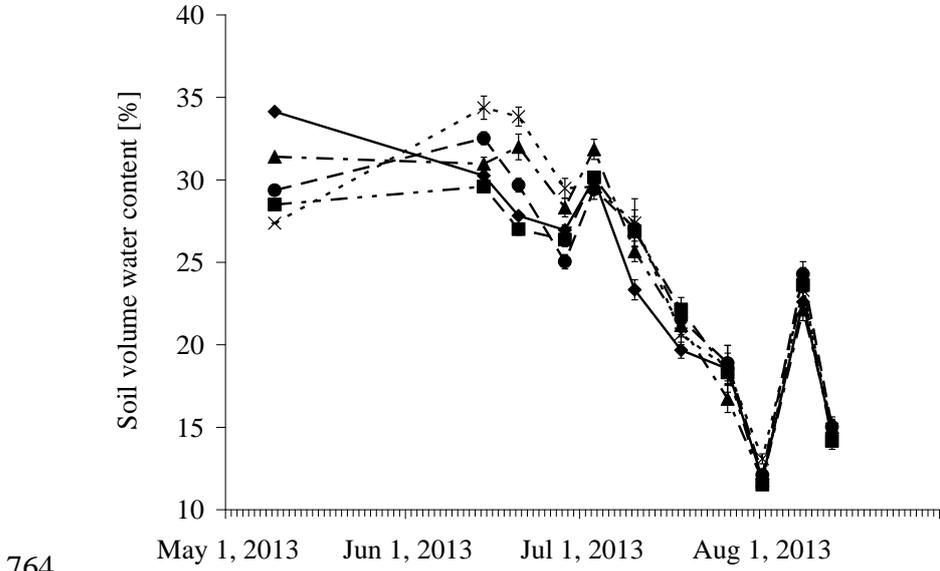
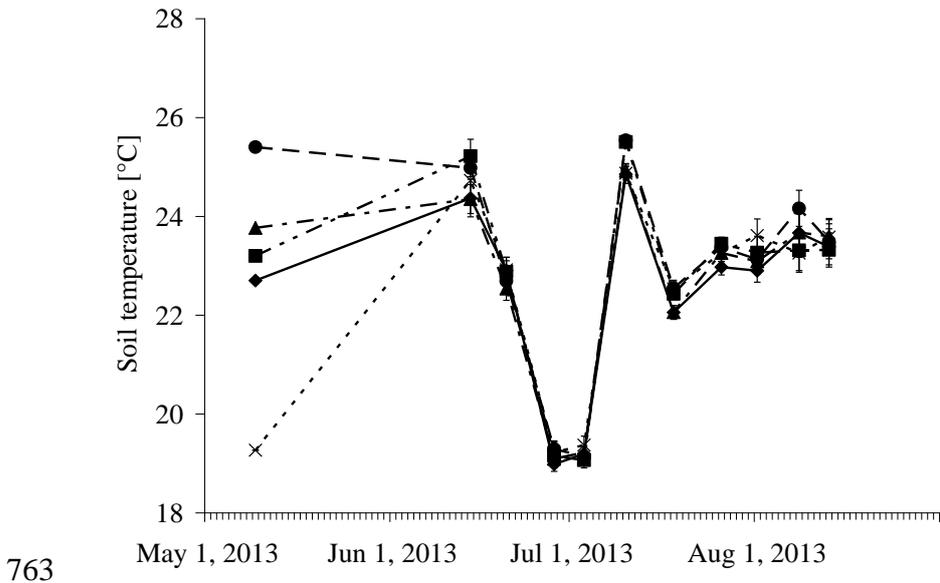
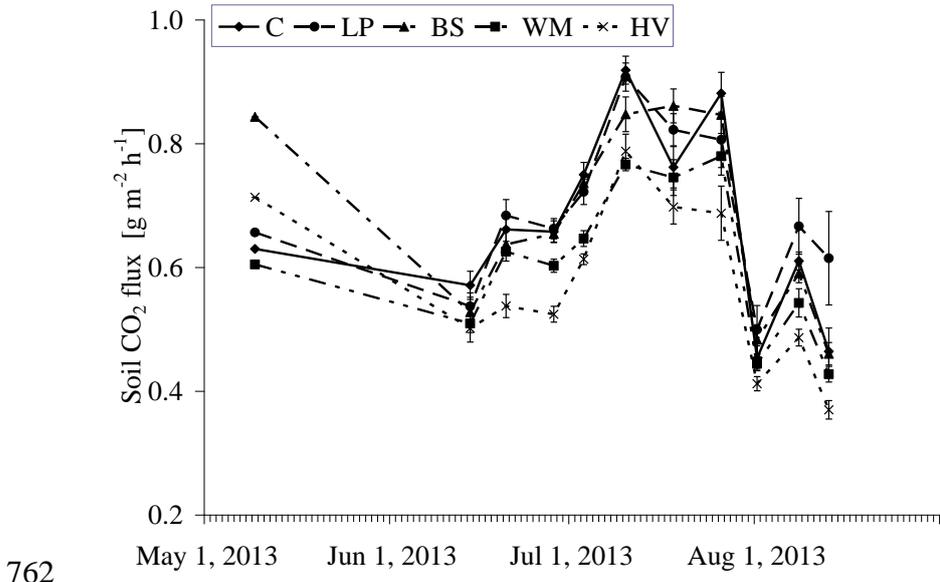


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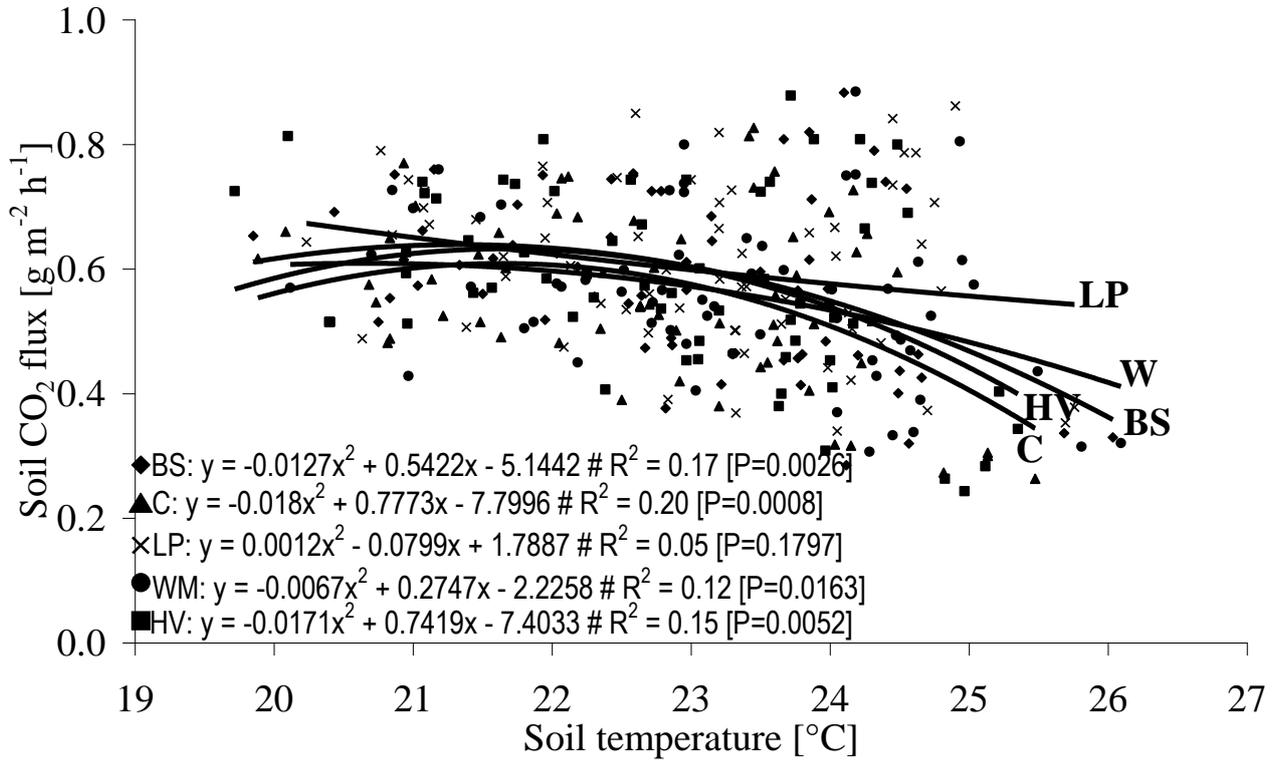


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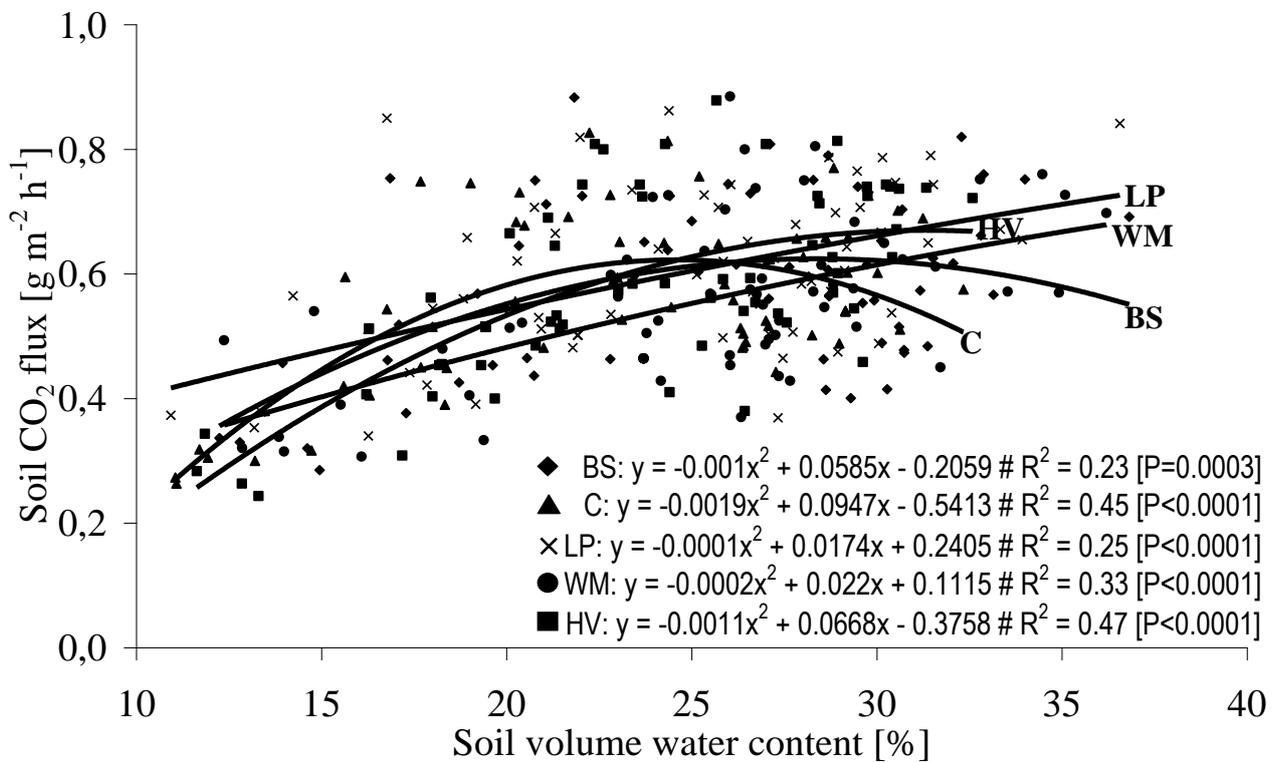
759 Figure 3 - Soil CO₂ flux, soil temperature and soil volume water content in the various mulching
 760 treatments, during the 2013 tomato cycle. Bars are standard errors (n = 18). BS: barley straw; C:
 761 conventional; HV: hairy vetch; LP: lacy phacelia; WM: white mustard.



766 Figure 4 – Soil CO₂ flux plotted against soil temperature and soil volume water content in the
 767 various mulching treatments. The data fit with second-order polynomial regression models. BS:
 768 barley straw; C: conventional; HV: hairy vetch; LP: lacy phacelia; WM: white mustard.

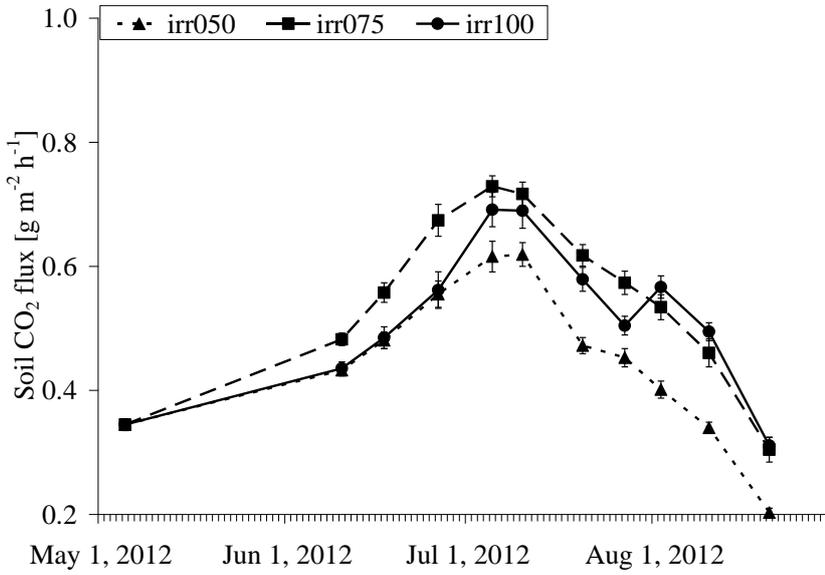


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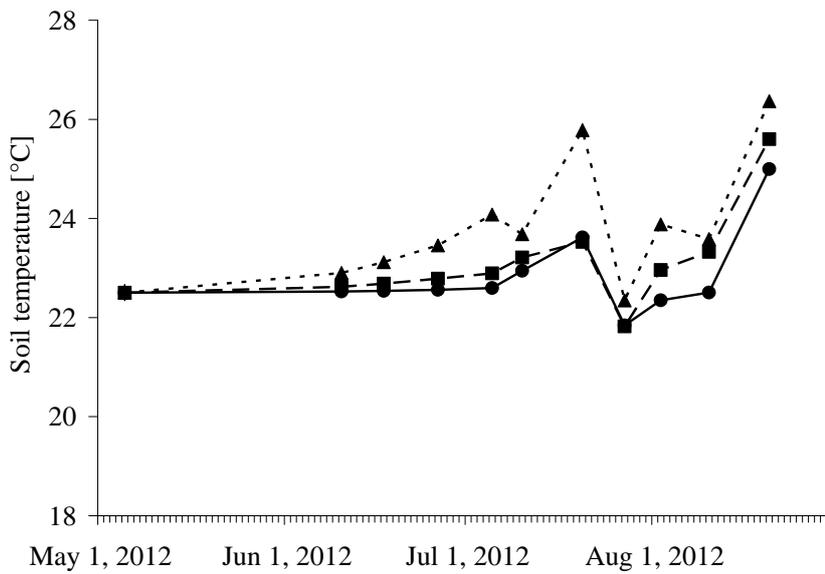


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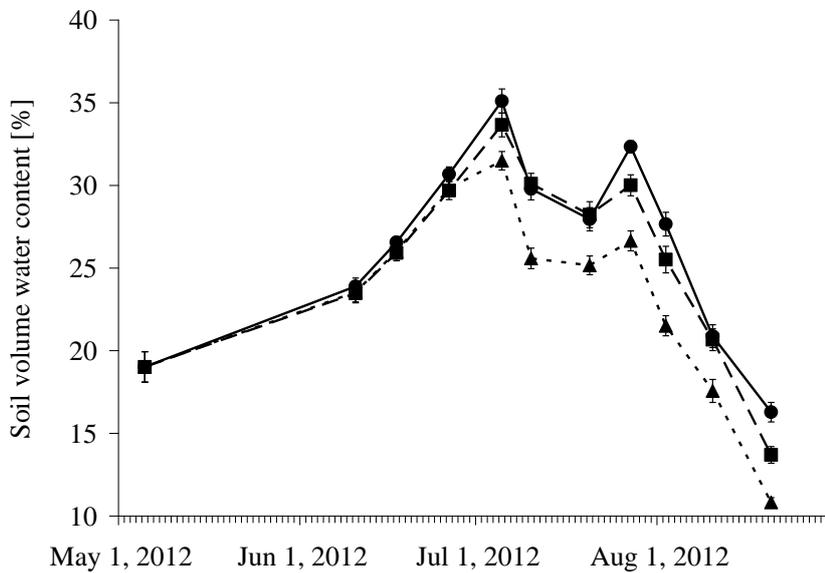
773 Figure 5 - Soil CO₂ flux, soil temperature and soil volume water content in the various irrigation
 774 levels, during the 2012 tomato cycle. Bars are standard errors (n = 30). irr50: irrigation 50%; irr75:
 775 irrigation 75%; irr100: irrigation 100%.



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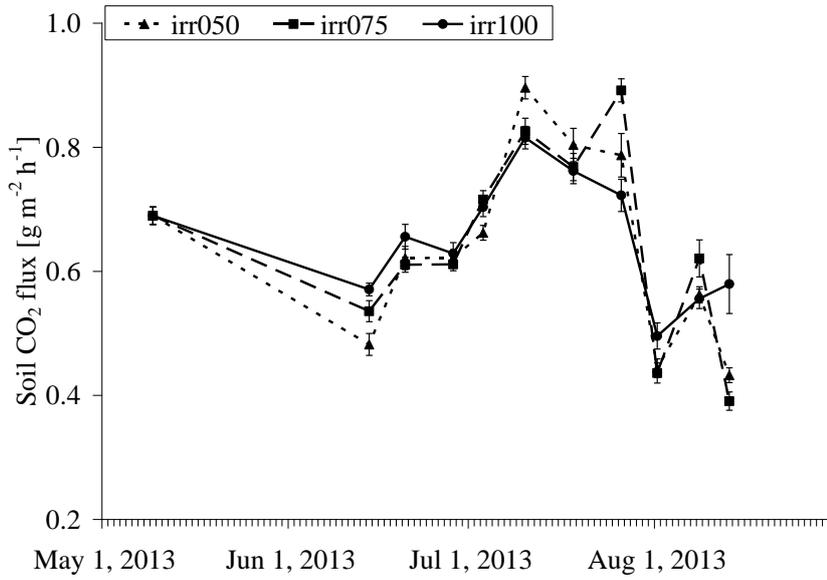


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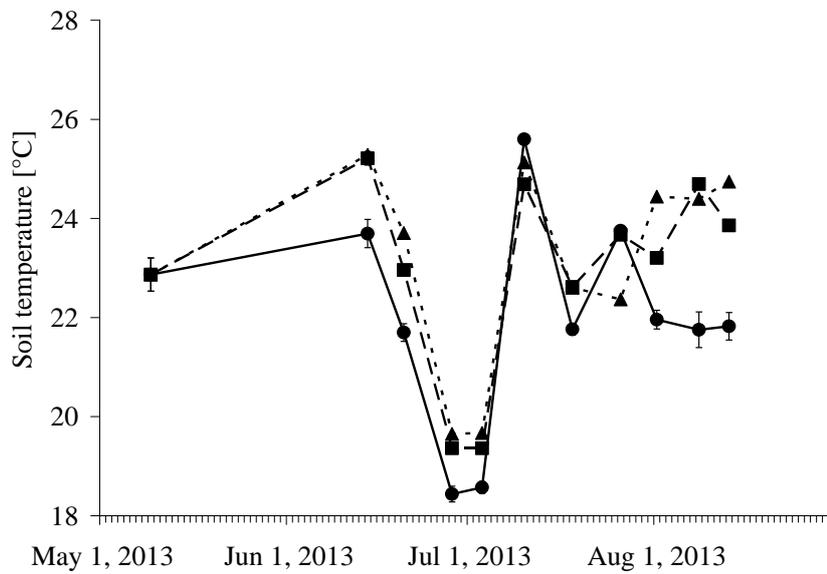


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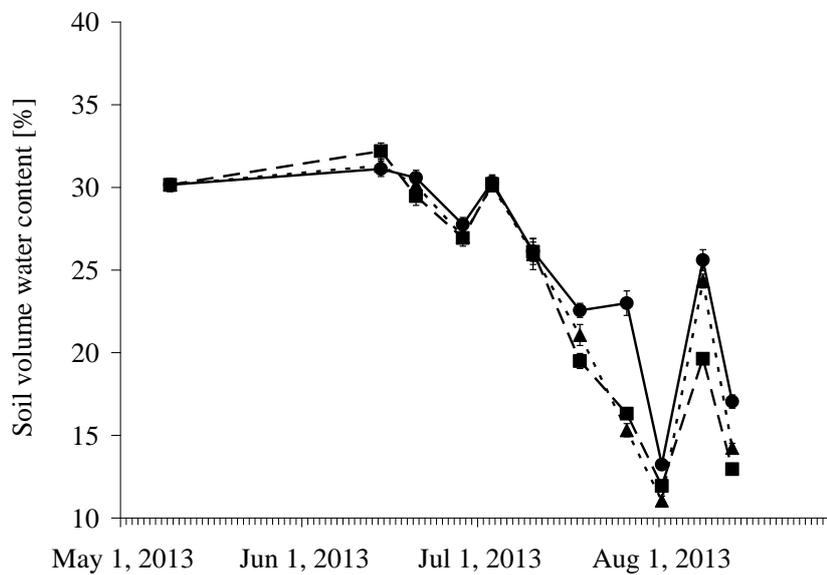
780 Figure 6 - Soil CO₂ flux, soil temperature and soil volume water content in the various irrigation
 781 levels, during the 2013 tomato cycle. Bars are standard errors (n = 30). irr50: irrigation 50%; irr75:
 782 irrigation 75%; irr100: irrigation 100%.



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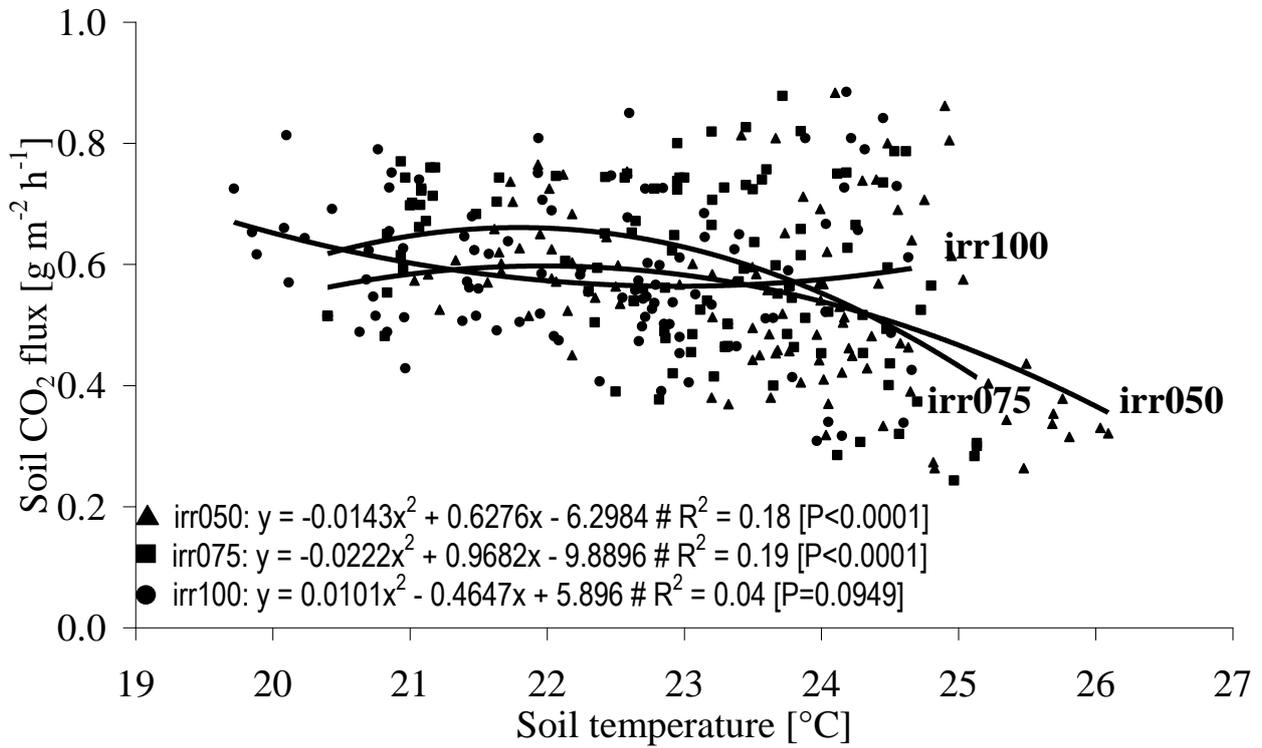


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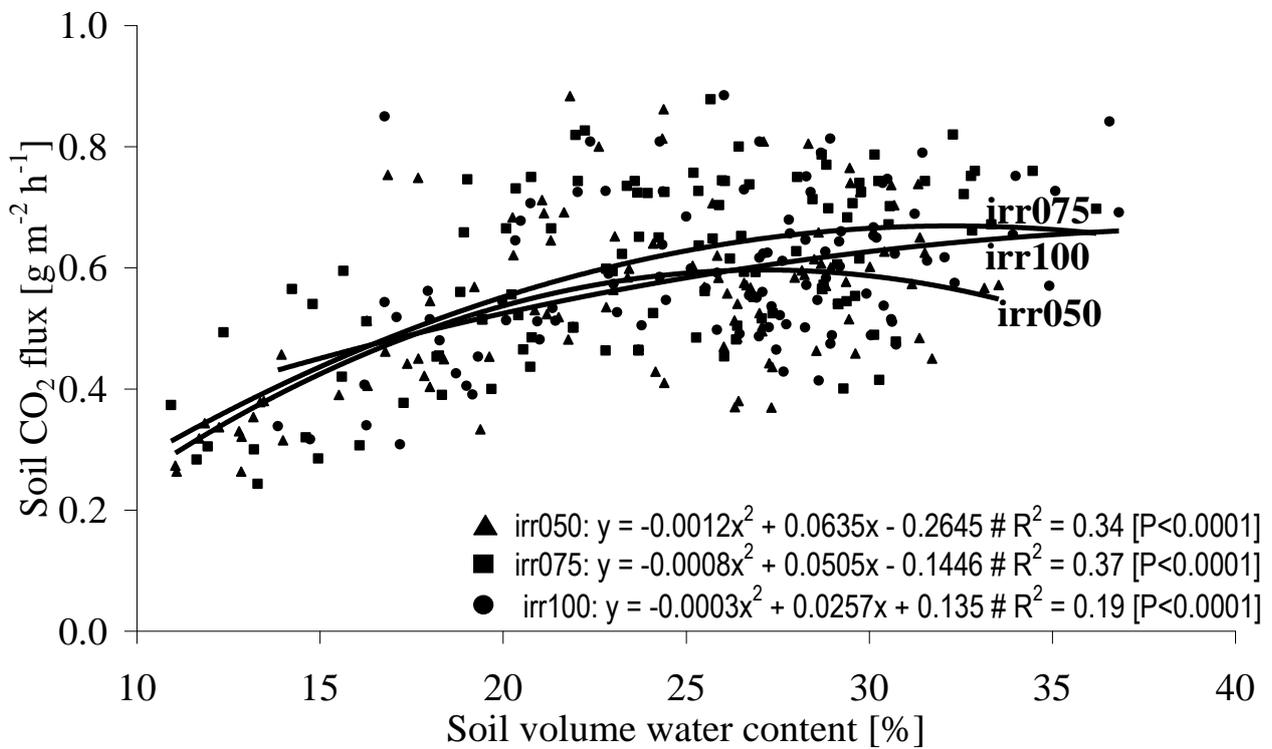


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787 Figure 7 – Soil CO₂ flux plotted against soil temperature and soil volume water content at various
 788 irrigation levels. The data fit with second-order polynomial regression models. irr50: irrigation
 789 50%; irr75: irrigation 75%; irr100: irrigation 100%.

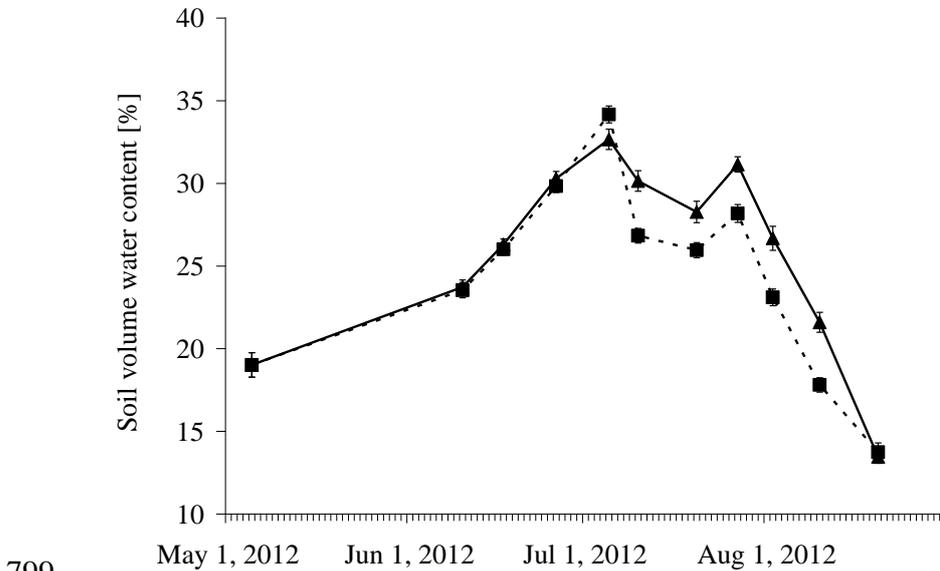
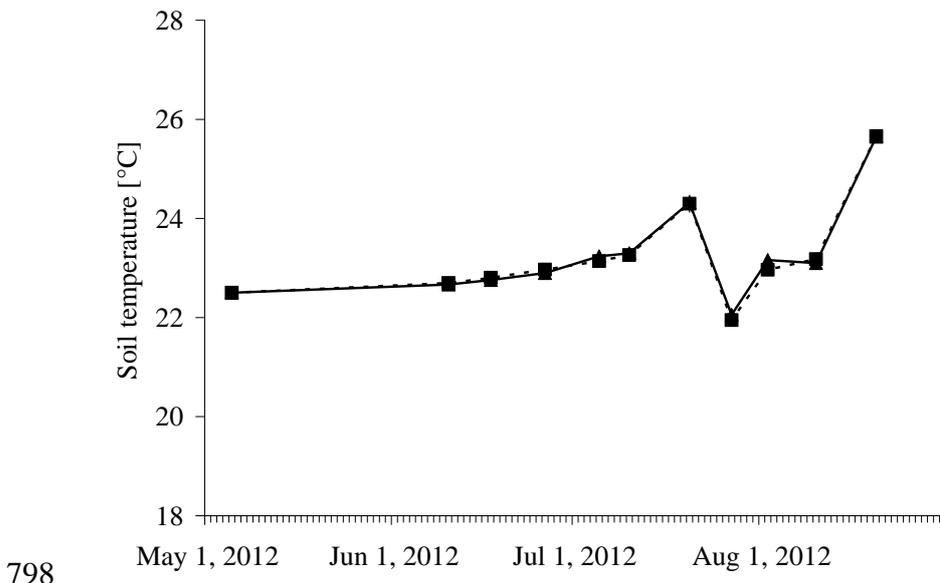
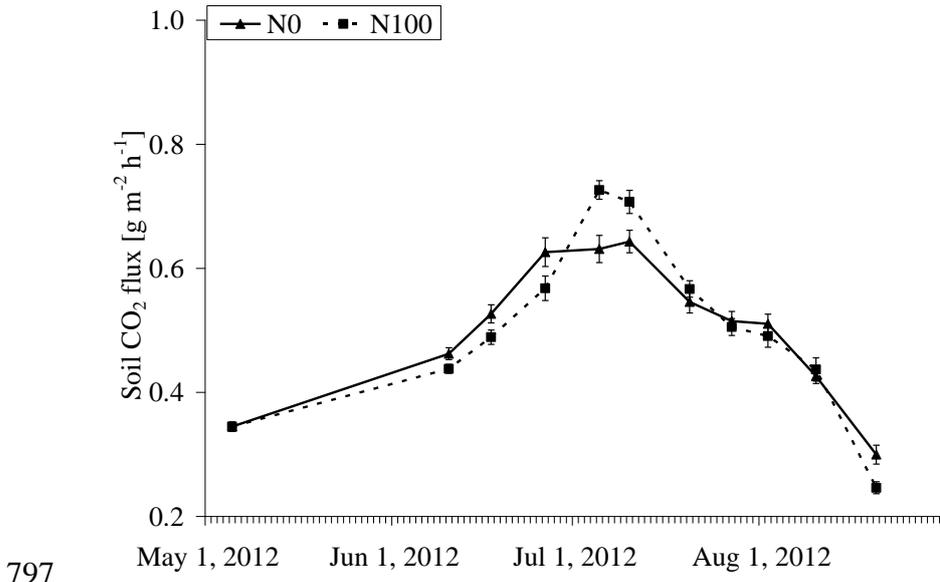


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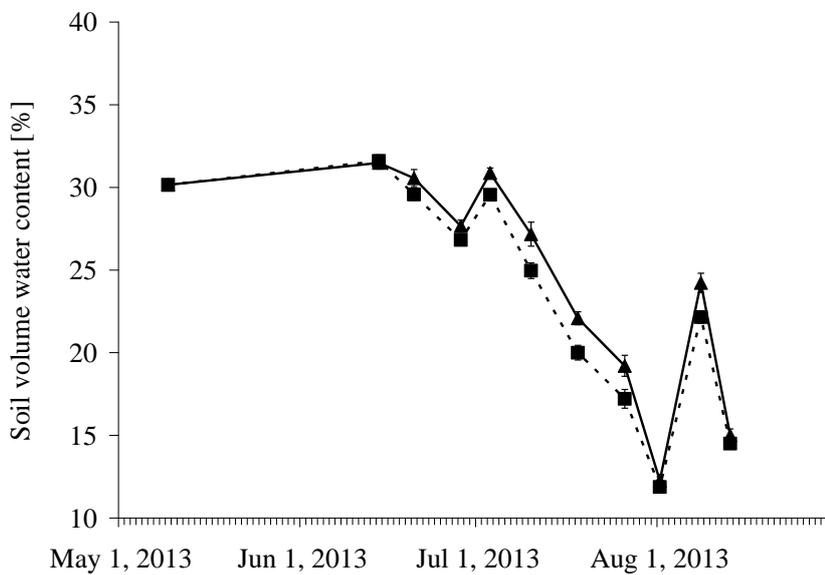
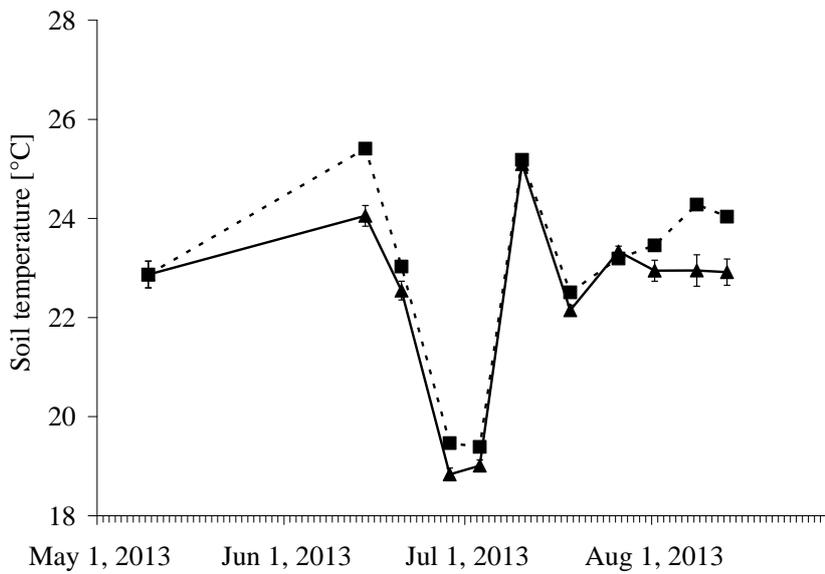
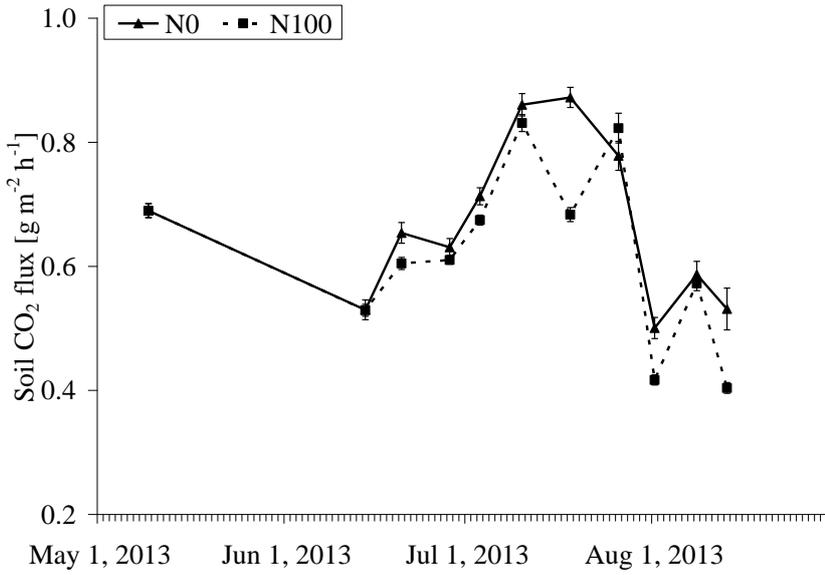


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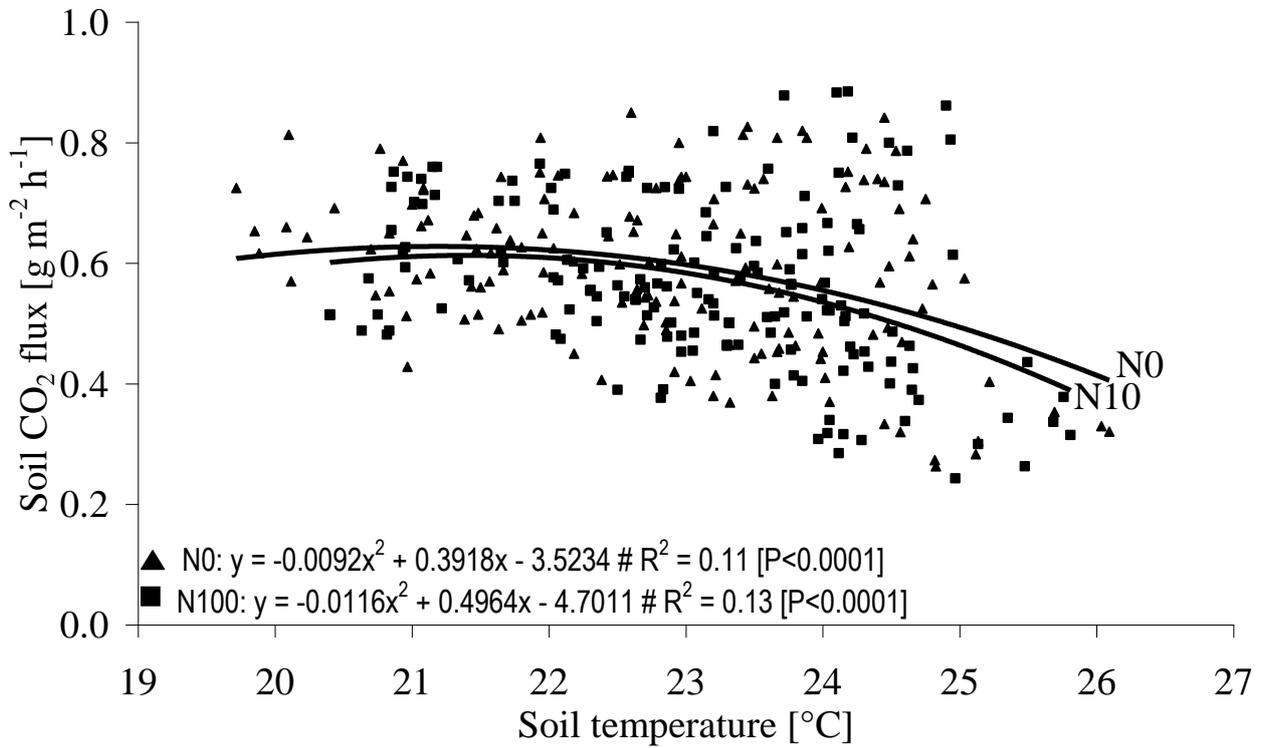
794 Figure 8 - Soil CO₂ flux, soil temperature and soil volume water content in the two fertilization
 795 levels, during the 2012 tomato cycle. Bars are standard errors (n = 45). N0: unfertilized; N100:
 796 fertilized.



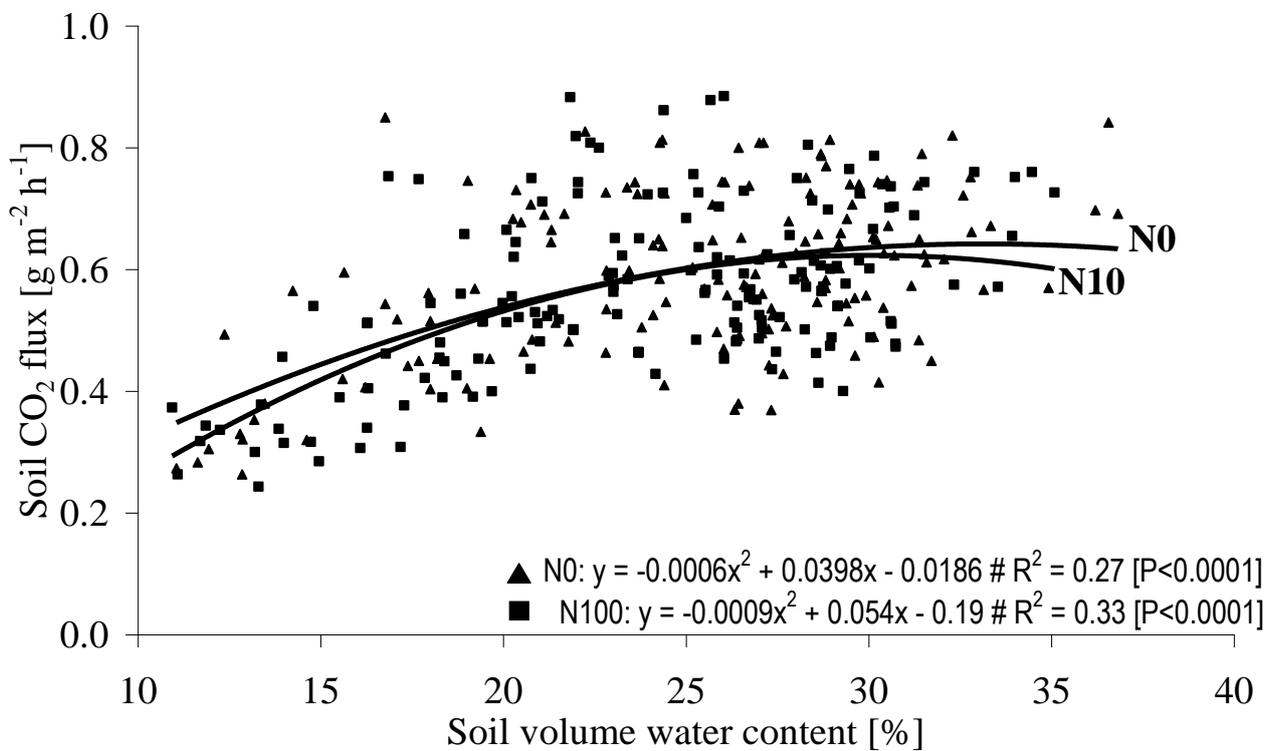
801 Figure 9 - Soil CO₂ flux, soil temperature and soil volume water content in the two fertilization
 802 levels, during the 2013 tomato cycle. Bars are standard errors (n = 45). N0: unfertilized; N100:
 803 fertilized.



808 Figure 10 – Soil CO₂ flux plotted against soil temperature and soil volume water content in the two
809 different fertilization levels. The data fit with second-order polynomial regression models. N0:
810 unfertilized; N100: fertilized.



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815 Table 1 - The interaction effect of the mulching x fertilization and mulching x irrigation on soil
 816 TOC and TON at tomato harvesting in the two experimental years. Values belonging to the same
 817 parameter and year with different letters in rows for mulching effects (upper case letter), and in
 818 columns for fertilization and irrigation effect (lower case letter) are statistically different according
 819 to LSD (0.05).

Mulching		Soil TOC (%)		Soil TON (%)	
		N-0	N-100	N-0	N-100
2012	conventional	1.122 Abc	1.070 Ae	0.111 Bab	0.140 Aa
	lacy phacelia	1.171 Bab	1.294 Aa	0.112 Bab	0.143 Aa
	barley straw	1.162 Aab	1.188 Abc	0.116 Bab	0.128 Ab
	white mustard	1.074 Bc	1.155 Acd	0.106 Bb	0.128 Ab
	hairy vetch	1.204 Aa	1.113 Bde	0.121 Aa	0.116 Ac
2013	conventional	1.588 Ab	1.531 Ab	0.143 Ae	0.139 Ab
	lacy phacelia	1.695 Aa	1.716 Aa	0.161 Abd	0.163 Aa
	barley straw	1.701 Aa	1.719 Aa	0.160 Acd	0.154 Aa
	white mustard	1.739 Aa	1.644 Ba	0.197 Aa	0.165 Ba
	hairy vetch	1.702 Aa	1.703 Aa	0.154 Bde	0.165 Aa
		irr50	irr100	irr50	irr100
2012	conventional	1.166 Aab	1.026 Be	0.146 Aa	0.106 Bd
	lacy phacelia	1.191 Ba	1.274 Aa	0.127 Abc	0.127 Aab
	barley straw	1.155 Aab	1.196 Abc	0.115 Be	0.128 Aab
	white mustard	1.117 Ab	1.112 Ad	0.116 Ade	0.118 Ac
	hairy vetch	1.151 Aab	1.166 Acd	0.119 Ace	0.119 Abc
2013	conventional	1.580 Ac	1.539 Ad	0.145 Ac	0.138 Ad
	lacy phacelia	1.711 Aa	1.699 Abc	0.164 Aab	0.159 Ac
	barley straw	1.640 Bac	1.780 Aa	0.152 Abc	0.161 Abc
	white mustard	1.713 Aa	1.670 Ac	0.168 Aa	0.194 Ba
	hairy vetch	1.622 Bbc	1.783 Aa	0.161 Aab	0.159 Ac

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824 Table 2 - The cover crops aboveground biomass and relative C and N content and the effect of the
 825 mulching, irrigation, and fertilization on the tomato yield in the two experimental years. Values
 826 belonging to the same parameter and treatment with different letters in columns are statistically
 827 different according to LSD (0.05).
 828

	Cover crops aboveground biomass (g m ⁻² of dm)	C in the aboveground biomass (%)	N in the aboveground biomass (%)	Tomato yield (kg m ⁻² of FM)	
2012	conventional			5.23 b	
	lacy phacelia	439 b	39.3 b	0.8 b	4.26 cd
	barley straw				4.02 de
	white mustard	365 b	43.9 a	1.3 b	3.81 e
	hairy vetch	785 a	42.3 a	2.7 a	5.94 a
	irr50				3.02 c
	irr75				5.01 b
	irr100				5.92 a
	N-0				3.85 b
	N-100				5.46 a
2013	conventional			3.56 b	
	lacy phacelia	525 a	39.5 b	1.4 b	2.90 d
	barley straw				3.28 c
	white mustard	365 b	41.5 a	1.5 b	2.44 e
	hairy vetch	564 a	41.4 a	3.8 a	5.44 a
	irr50				2.31 c
	irr75				3.86 b
	irr100				4.41 a
	N-0				2.75 b
	N-100				4.31 a

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831 Table 3 - The effect of the mulching, irrigation, and fertilization on the C balance in the two
 832 experimental years. Values belonging to the same parameter and treatment with different letters in
 833 columns are statistically different according to LSD (0.05).
 834

		C input by biomasses	C output by CO ₂ fluxes	input/output rate
		Mg C ha ⁻¹		
2012	conventional	2.15 e	3.15 a	0.73 d
	lacy phacelia	4.37 bd	3.43 a	1.35 bd
	barley straw	3.69 d	3.19 a	1.21 cd
	white mustard	4.13 cd	3.34 a	1.46 ad
	hairy vetch	7.8 a	3.88 a	2.1 a
	irr50	4.13 b	3.09 b	1.43 a
	irr75	4.58 a	3.66 a	1.32 a
	irr100	4.59 a	3.45 ab	1.36 a
	N-0	4.11 b	3.41 a	1.28 b
	N-100	4.75 a	3.39 a	1.46 a
2013	conventional	1.76 e	4.17 a	0.43 e
	lacy phacelia	4.38 bc	4.18 a	1.09 bd
	barley straw	3.47 d	4.35 a	0.83 de
	white mustard	3.96 cd	3.81 a	1.05 cd
	hairy vetch	6.73 a	3.73 a	1.87 a
	irr50	3.78 b	3.99 a	0.99 b
	irr75	4.11 ab	4.08 a	1.05 ab
	irr100	4.3 a	4.08 a	1.12 a
	N-0	3.85 b	4.14 a	0.98 b
	N-100	4.27 a	3.95 a	1.13 a

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