

Evaluation of the potential of the current and forthcoming multispectral and hyperspectral imagers to estimate soil texture and organic carbon

Castaldi Fabio¹, Palombo Angelo², Santini Federico², Pascucci Simone², Pignatti Stefano², Casa Raffaele¹

¹ *Department of Agricultural and Forestry sciNcEs (DAFNE), Università degli Studi della Tuscia, Via San Camillo de Lellis, 01100 Viterbo, Italy.*

² *Consiglio Nazionale delle Ricerche- Institute of Methodologies for Environmental Analysis (C.N.R. – IMAA), Via del Fosso del Cavaliere, 100, 00133 Roma, Italy.*

Correspondence: F.Castaldi, castaldi@unitus.it

Abstract

In this study the capabilities of seven multispectral and hyperspectral satellite imagers to estimate soil variables (clay, sand, silt and organic carbon content) were investigated using data from soil spectral libraries. Four current (EO-1 ALI and Hyperion, Landsat 8 OLI, Sentinel-2 MSI) and three forthcoming (EnMAP, PRISMA and HypSIRI) satellite imagers were compared. To this aim, two soil spectra datasets that simulated each imager were obtained: (i) resampled spectra according to the specific spectral response and resolution of each satellite imager and (ii) resampled spectra with declared or actual noise (radiometric and atmospheric) added. Compared with those using full spectral resolution data, the accuracy of Partial Least Square Regression (PLSR) predictive models generally decreased when using resampled spectra. In the absence of noise, the performances of hyperspectral imagers, in terms of Ratio of Performance to Interquartile Range (RPIQ), were generally significantly better than those of multispectral imagers. For instance the best RPIQ for sand estimation was obtained using EnMAP simulated data (2.56), whereas the outcomes gained using multispectral imagers varied from 1.56 and 2.28. The addition of noise to the simulated spectra brought about a decrease of statistical accuracy in all estimation models, especially for

26 Hyperion data. Although the addition of noise reduced the performance differences between
27 multispectral and hyperspectral imagers, the forthcoming hyperspectral imagers nonetheless
28 provided the best RPIQ values for clay (2.16–2.33), sand (2.10–2.17), silt (2.77–2.85) and
29 organic carbon (2.48–2.51) estimation. To better understand the impact of spectral resolution
30 and signal to noise ratio (SNR) on the estimation of soil variables, PLSR models were applied
31 to resampled and simulated spectra, iteratively increasing the bandwidth to: 10, 20, 40, 80 and
32 160 nm. Results showed that, for a bandwidth of 40 nm, i.e., a spectral resolution lower than
33 that of current and forthcoming imagers, the estimation accuracy was very similar to that
34 obtained with a higher spectral resolution.

35 Forthcoming hyperspectral imagers will therefore improve the accuracy of soil variables
36 estimation from bare soil imagery with respect to the results achievable by current
37 hyperspectral and multispectral imagers. This work provides useful indications for the future
38 application of satellite data in Soil Science, and for the estimation of the most important soil
39 variables by next generation satellite imagers.

40 *Keywords: imaging spectroscopy, sand, clay, spectral library, SNR, PLSR, PRISMA, EnMAP,*
41 *HyspIRI*

42

43 **1. Introduction**

44

45 Understanding variability of soil properties between and within agricultural fields allows for
46 more efficient use of resources, improving agronomic and environmental management. The
47 qualitative information included in existing soil maps is often insufficient for site-specific
48 management strategies concerning water, fertilizers, herbicides or harvest. For these purposes,
49 the quantitative estimation of soil properties (e.g., soil texture, organic carbon, nitrogen, and

50 soil moisture) over the field is necessary. With the exception of a few regions in highly
51 developed agricultural environments, this kind of information is rarely available to land
52 managers.

53 Remote sensing data can be used to obtain, in a very cost effective way, qualitative and
54 quantitative information about soil variables and soil classification (Mulder et al., 2011). In
55 cultivated soils, due to repeated tillage operations, soil properties are usually quite uniform
56 over the tilled layer; therefore they can be estimated from the bare soil surface reflectance
57 (Casa et al., 2013b).

58 Quantitative estimation of soil variables using bare soil imagery acquired from multispectral
59 remote imagers is, however, hampered by inadequate spectral resolution, particularly by the
60 absence of narrow bands in the short wave infra-red (SWIR) region (1100–2400 nm) (i.e., the
61 spectral region more affected by the soil chromophores). For these reasons, multispectral
62 satellite data are mainly used for qualitative assessments, such as the classification of areas
63 with different soil textures (e.g., Demattê et al., 2009; Odeh & McBratney, 2000; Zhai et al.,
64 2006). Recent studies obtained a sufficient degree of accuracy in the quantitative estimation
65 of silt (Wu et al., 2015) or clay (Castaldi et al., 2014) using, respectively, BJ-1 and Advanced
66 Land Imager (ALI) satellite imagers. It should be noted that both the multispectral ALI
67 imager on board the NASA EO-1 satellite and the Operational Land Imager (OLI) imager on
68 board the Landsat 8 satellite, have bands in the SWIR region, which can be exploited for soil
69 properties estimation. Sentinel-2, which was successfully launched in June 2015, has a
70 Multispectral Imager (MSI) with a band in the SWIR region, between 2100 and 2280 nm and
71 centred at 2190 nm, with a Ground Sampling Distance (GSD) of 20 m.

72 Hyperspectral imagers, that measure spectral radiance for hundreds of narrow bands are more
73 attractive than multispectral imagers for soil spectroscopy purposes. Furthermore, over the
74 last decade, analysis of data obtained by optical remote sensing techniques such as soil

75 spectroscopy and hyperspectral imagery has proven to be an effective way to characterize and
76 monitor surface soil variables, that even allow soil erosion processes to be detected (Gomez et
77 al., 2015; Stevens et al., 2013; Ben-Dor et al., 2009; Lagacherie et al., 2008; Gomez, et al.,
78 2008). The higher spectral resolution provided by hyperspectral sensors could, in principle,
79 allow even more accurate quantitative estimates at the field scale, compared with those
80 obtainable from the existing multispectral imagers (Mulder et al., 2011).

81 Only two hyperspectral satellite imagers with a sufficient spatial resolution (i.e., $GSD \leq 30$ m)
82 are currently available for soil applications: Hyperion on board of the NASA EO-1 platform
83 and Compact High Resolution Imaging Spectrometer (CHRIS) on the European Space
84 Agency's PROBA platform. Both these sensors have considerable limitations in quantitative
85 soil estimation applications. Hyperion's data are hampered by the very low signal to noise
86 ratio (SNR) in the SWIR region, in particular around 2200 nm, where the spectral features of
87 clay minerals are located (Castaldi et al., 2014). The difficulties in estimating soil variables
88 from CHRIS-PROBA are due to its restricted spectral range (415–1050 nm) that lacks bands
89 in the SWIR region (Casa et al., 2013a). For these reasons, the use of satellite hyperspectral
90 data in quantitative soil estimation is still challenging and consequently the number of
91 published studies in which this type of data are used is still small (Casa et al., 2013a; Casa et
92 al., 2013b; Castaldi et al., 2014; Gomez et al., 2008; Zhang et al., 2013).

93 In the near future at least four satellites equipped with hyperspectral imagers are due to be
94 launched: the Japanese Hyperspectral Imager Suite (HISUI) in 2017 (Tanii et al., 2012); the
95 Italian PRecursores IperSpettrale della Missione Applicativa (PRISMA) in 2017 (Pignatti et
96 al., 2012); the German Environmental Mapping and Analysis Program (EnMap) in 2018
97 (Richter et al. 2012); the China Commercial Remote-sensing Satellite System (CCRSS) after
98 2018; and the U.S. NASA Hyperspectral Infrared Imager (HyspIRI) in 2021 (Houborg et al.
99 2012). The Spaceborn Hyperspectral Applicative Land and Ocean Mission (SHALOM) — a

100 joint mission by the Israel Space Agency (ISA) and the Italian Space Agency (ASI) — will
101 also develop a hyperspectral imager with 241 bands between 400 and 2500 nm and a spectral
102 resolution of about 10 nm (Ben-Dor et al., 2013). A new hyperspectral imager, HYPerspectral
103 X Imagery (HYPXIM) is also under study by the French space agency (CNES) (Michel et al.,
104 2011). Forthcoming hyperspectral imagers will have numerous narrow bands in the SWIR
105 spectral region, which will presumably permit accurate estimation of soil variables; however,
106 the soil properties estimation accuracy of these imagers will depend on their SNR, particularly
107 around 2200 nm. Hyperspectral imagers generally have a lower SNR than multispectral ones
108 as a result of the reduced energy collected by the sensor in narrow spectral bands. This effect,
109 coupled with the low solar irradiance in the SWIR region, produces a consistent decrease of
110 the SNR. For example, Castaldi et al. (2014) compared the soil estimation capabilities of two
111 sensors mounted on EO-1 satellite using both Hyperion and ALI data. The authors did not
112 observe any apparent advantages when using hyperspectral (Hyperion) instead of
113 multispectral (ALI) data. This was explained by the low SNR of Hyperion at the wavelengths
114 of characteristic spectral features of clay minerals.

115 This study aims to evaluate the performances of current and forthcoming multispectral and
116 hyperspectral imagers for the quantitative retrieval of soil texture and Soil Organic Carbon
117 (SOC). To this end we compare the estimation accuracy for soil texture (clay, sand and silt)
118 and SOC using spectra acquired under laboratory conditions, which were resampled
119 according to spectral and noise characteristics of four current (ALI, Landsat 8, Sentinel-2 and
120 Hyperion) and three forthcoming (EnMAP, PRISMA, HypsIRI) satellite imagers. To our
121 knowledge, no previous reports have specifically compared the capability of these imagers for
122 soil texture and SOC estimation.

123 2. Materials and Methods

124 2.1. Soil spectral libraries

125 2.1.1. PONMAC library

126

127 A soil spectral library consisting of 166 samples was assembled by pooling together data from
128 soil samplings carried out in two cropland areas in Central and Southern Italy. Samplings
129 were carried out in Pontecagnano (PON; Southern Italy, near to Salerno) and Maccarese
130 (MAC; Central Italy, near to Rome) and a pooled dataset obtained from the union of PON
131 (Pascucci et al., 2014) and MAC (Casa et al., 2013b), hereafter referred to as PONMAC. The
132 soils of the MAC area are classified as Cutanic Luvisol (FAO-ISRIC-ISSS, 1998), with soil
133 parent materials of flat inshore deposits (Pleistocene), while the soils of PON area originated
134 from travertine sediments characterized by sandy gravel layers with tuffaceous intercalation in
135 the upper parts (Pleistocene–Holocene). In both areas, soil sampling was carried out using a
136 gouge auger at 0–10 cm depth in Pontecagnano and at 0–30 cm depth in Maccarese. Soil
137 samples were air dried and passed through a 2 mm sieve. For each sample we measured the
138 percentage of clay, sand and silt contents using the pipette method according to the United
139 States Department of Agriculture (USDA) system (Soil Survey Staff, 2014). Soil Organic
140 Content (%) was obtained using the Walkley-Black method for PON data and an elemental
141 analyzer (Flash EA1112, Thermo Electron Corporation, U.S.A.) according to the technique of
142 dry combustion analysis (ISO 10694, 1995) for MAC data. Although SOC measurements
143 were carried out using different methods for MAC and PON dataset, the results obtained from
144 Walkley-Black and dry combustion analysis were considered to be comparable within the
145 range of SOC values of the PONMAC dataset (Chen et al., 2015).

146 Soil textures of PONMAC samples are mainly composed of sandy clay loam, clay loam and
147 clay. These textural classes are among the most frequent in the croplands of Italy (Costantini
148 et al., 2012). The ranges of the relative contents (%) of clay, sand and silt in the PONMAC
149 dataset are quite wide (Table 1). The SOC content range is between 0.5 and 2.32%, with a
150 mean value of 1.25% (Table 1).

151 Soil samples were placed in Petri dishes and their spectral signatures were measured in a dark
152 lab in the visible-near infrared (VNIR) to SWIR optical domain (350-2500 nm, spectral
153 sampling of 1 nm) using an Analytical Spectral Devices (ASD) Field Spec Fr Pro
154 spectroradiometer (ASD Inc., Boulder, CO, USA; available at: <http://www.asdi.com>)
155 equipped with a contact probe containing a 7 W quartz-halogen lamp. Reflectance values
156 from 350 to 399 nm and from 2401 and 2500 nm were removed prior to any processing
157 because these spectral ranges are affected by noise.

158 The ASD measurements consisted of three spectral acquisitions at nadir rotating the sample
159 by 90° each time. Each replicate involved 50 spectral scans and after every set of 50, a
160 Spectralon panel (Labsphere, NH, USA) was measured before the sample was rotated. The
161 average measured radiance was then converted into bi-conical reflectance using the calibrated
162 Spectralon panel.

163

164 **2.1.2. LUCAS library**

165

166 In the framework of European Land/Use Cover Area frame Statistical Survey (LUCAS) about
167 20,000 topsoil samples were collected across Europe (Eurostat, 2009; Tóth et al., 2013). The
168 soil properties of the LUCAS dataset were analyzed using ISO standards methods. The
169 determination of particle size distribution was carried out using the sieving and sedimentation

170 method (ISO 11277. 1998), and SOC was estimated by dry combustion using an elemental
171 analyzer (ISO 10694:1995). Spectra were acquired on diffuse high-resolution reflectance for
172 all air-dried and sieved (< 2 mm) soil samples using an XDS Rapid Content Analyzer (FOSS
173 NIRSystems Inc., Laurel, MD, USA) spectroscope, measuring a continuous (spectral
174 resolution of 0.5 nm) reflectance spectrum in the VNIR (400–1300 nm) and SWIR (1300–
175 2500 nm) domains. Soil spectroscopy measurements were made following the protocol of the
176 Soil Spectroscopy Group (SPS, 2011).

177 In order to have a dataset comparable with PONMAC, we extracted from the LUCAS topsoil
178 dataset only the samples collected on cropland areas in Italy (LUCAS_C) and we resampled
179 the LUCAS spectra according to the spectral resolution of the PONMAC library (1 nm). The
180 713 samples thus extracted from the European LUCAS library are representative of all the
181 soil regions in the Soil Map of Italy (Costantini et al., 2012) and display a large variability in
182 clay, sand, silt and SOC contents (Table 1).

183 The PONMAC and LUCAS_C datasets cannot be fused into a single dataset however because
184 different spectral acquisition techniques and protocols were used to measure reflectance.

185

186 **2.2. Satellite sensors**

187

188 The performances of seven satellite imagers were compared. We took into consideration three
189 operating multispectral imagers (EO-1 ALI, Sentinel-2 MSI and LANDSAT 8 OLI), one
190 operating hyperspectral imager (EO-1 Hyperion) and three forthcoming hyperspectral imagers
191 (EnMAP, PRISMA and HypsIRI). The main technical specifications of the multispectral and
192 hyperspectral imagers are summarized, respectively, in Table 2 and Table 3. In particular, the
193 last two columns of both tables show the SNR at different wavelengths and the related signal

194 intensities (Pignatti et al., 2013; Baillarin et al., 2012; Markham et al. 2012; Sang et al., 2008;
195 Green et al., 2008; Mendenhall et al., 2001; Willoughby et al., 1996).

196

197 **2.3. Resampled and simulated data**

198

199 The LUCAS_C and PONMAC spectral libraries were used to produce two sets of synthetic
200 spectra. A first set of spectra was obtained by resampling the library spectra, using
201 convolution procedures, to the specific spectral response and resolution of the various sensors.
202 This process was carried out using the spectral response function for the multispectral
203 imagers, while for hyperspectral imagers the spectral response functions were not available
204 and were subsequently approximated using Gaussian functions. The spectra obtained in this
205 way are hereafter referred to as "resampled spectra".

206 In order to simulate the satellite data more realistically, an additional set of data was obtained
207 by adding spectral noise to the synthetic spectra previously generated. For this purpose, we
208 firstly applied a direct process that allowed us to simulate the spectral top of the atmosphere
209 (TOA) at-sensor radiance, under specific illumination and acquisition geometries. The direct
210 process can be represented by the equation (Vermote et al., 1997; Kaufmann et al., 1997):

$$211 \quad L = \frac{A\rho}{1 - \rho S} + L_a \quad [1]$$

212 where L is the at-sensor radiance, L_a is the solar radiation scattered back to the sensor without
213 reaching the soil (path radiance), ρ is the target reflectance, A is a parameter depending on the
214 atmosphere alone (not on the illumination and acquisition geometries), and S is the spherical
215 albedo of the atmosphere. In our simulations ρ was taken from the spectral libraries used for
216 this study, while A , S and L_a were simulated by MODTRAN 4 code, using a standard mid-

217 latitude summer model to characterize the atmospheric column. The illumination geometry
218 corresponds to the geographic coordinates of the center of Rome (41°53'35.94"N;
219 12°28'58.57"E) on the summer solstice at noon, local time. Once obtained, a spectral noise,
220 consistent with the sensor radiometric characteristics and acquisition parameters, was added
221 to the at-sensor TOA radiances.

222 In order to properly represent the noise dependence from the spectral radiance intensity
223 reaching the sensor, more information than that shown in Tables 2 and 3 is needed. For
224 HypsIRI and for all the multispectral imagers, with the exception of Sentinel-2, we found in
225 the literature the spectral SNR at different radiance magnitudes. Since the covered intensity
226 range was sufficiently wide, we could interpolate the SNR, thus obtaining look-up tables to
227 identify the at-sensor spectral SNR at any radiance intensity. To fill in the missing
228 information, we approached the research and technical teams of PRISMA and EnMAP for the
229 spectral SNR at four different radiance levels, and were provided spectral information
230 analogous to that already published for HypsIRI shown in Figure 1 (Green et al., 2008).
231 Instead, for Hyperion and Sentinel-2 we were only able to obtain a unique spectral SNR level
232 of a reference radiance, therefore we used a generic noise model to extrapolate the SNR at
233 other radiance intensities (Chang, 2007):

$$234 \quad SNR = SNR_{ref} \sqrt{\frac{L}{L_{ref}}} \quad [2]$$

235 where, SNR_{ref} and L_{ref} are the nominal SNR and the related incident radiance, respectively. In
236 this way, we managed to extrapolate the proper SNR at different illumination geometries and
237 intensities.

238 After the conversion of the SNR to Noise equivalent in Δ Radiance (Ne Δ R), the noise to be
239 added to Equation (3) was computed by applying Gaussian statistics. By using the described

240 direct process (i.e., Equation (1)), we simulated the TOA radiance spectra (noise included)
241 that each sensor would acquire for a set of different targets (reflectance spectra) at a specific
242 time and day of the year.

243 Finally, the inverted process allowed us to calculate the top of the canopy (TOC) reflectance
244 spectra as shown in Equation (3).

$$245 \quad \rho = \frac{1}{A} [L - L_a - Gauss(0, Ne\Delta R)](1 - \rho S) \quad [3]$$

246 In this way, for each original reflectance spectrum, we obtained a set of spectra consistent
247 with the spectral and radiometric characteristics of the different analyzed satellite imagers. In
248 other words, we simulated the spectra that would be obtained by the remote imagers after the
249 application of radiometric and atmospheric corrections under a given illumination condition.
250 These spectra are used to evaluate the potentials of the various imagers in the estimation of
251 soil variables and are hereafter referred to as "simulated spectra". Figure 2 displays two
252 Hyperion simulated reflectance spectra with (dashed line) and without (solid line) noise as
253 obtained using a spectrum belonging to the PONMAC spectral library.

254 In order to investigate the issue of the low SNR of hyperspectral sensors, due to the reduced
255 energy collected by the sensors in narrow spectral bands, the effect of broadening the
256 bandwidths on the estimation accuracy of soil variables was examined. Synthetic (both
257 resampled and simulated) spectra were iteratively resampled, doubling the bandwidth at each
258 step. In Figure 3 five spectra obtained by resampling a simulated HySPIRI spectrum
259 (PONMAC library) are depicted with five different bandwidths: 10, 20, 40, 80, and 160 nm.
260 From Figure 3 it is evident that the bandwidth increase determines the progressive smoothing
261 of some spectral features.

262

2.4. Soil variables estimation

263

264

265 We applied the Partial Least Square Regression (PLSR) technique (Wold et al., 2001) to
266 estimate soil variables from three types of data: (i) full laboratory spectra (400–2400 nm;
267 spectral resolution 1 nm); (ii) resampled spectra according to bands and spectral resolution of
268 each satellite’s imager; (iii) simulated spectra that were obtained by adding declared or actual
269 sensors’ noise, as well as atmospheric effects, to the resampled spectra.

270 In order to detect outliers, we used lower and upper limits of the distribution of each variable.
271 Limits were set at 1.5 times the inter-quartile range (IQ) and observations outside the upper
272 and lower value were considered outliers and eliminated from the following computations
273 (Tukey, 1977). The IQ is the difference between the values below which 75% (Q3) and 25%
274 (Q1) of the samples occur ($IQ = Q3 - Q1$). The number of samples removed as a result of this
275 process is reported in Table 1. Moreover, skewed data (skewness $>|0.5|$) were transformed
276 using the square root or logarithm of the variable (Table 1) to approximate a Gaussian
277 distribution.

278 Before applying the PLSR method to the simulated spectra, the noisy bands falling into the
279 absorption ranges of the atmospheric gases were removed. In Table 4 the removed spectral
280 bands for each sensor are shown.

281 The PLSR is a multivariate regression technique widely used in soil spectroscopy
282 applications. PLSR provides an optimal linear model that allows for the reduction of the
283 number of correlated predictors (spectral bands) and transforms them to a restricted number of
284 uncorrelated components (PLSR components), which have the best relationship with the
285 dependent variables (in this case clay, sand, silt and SOC). In order to detect the main spectral
286 regions that affect soil variables estimation, we calculated the variance importance in

287 projection (VIP) values. The VIP value is a weighted sum of squares of the PLS weights.
288 Weights are calculated from the amount of variance of dependent variables of each PLS
289 component employed in the model (Wold et al., 2001). Predictors with VIP values greater
290 than one are considered significant for the PLSR models because the average of VIP scores is
291 equal to one (Chong & Jun, 2005).

292 Different PLSR models were preliminarily tested, using different combinations of PLSR
293 components, spectral transformations and smoothing to obtain the most accurate prediction of
294 soil variables in terms of root mean square error (RMSE). For these purposes the spectra were
295 converted into absorbance (A) and/or first derivative (D) and the Savitzky-Golay (SG)
296 smoothing filter was applied. The Savitzky-Golay filtering performs noise reduction and
297 enhances small spectral differences; whereas derivative transformation of the spectra is useful
298 for separating out peaks of overlapping bands and to reduce noise.

299 Each dataset was randomly split into calibration and validation subsets: 70% for calibration
300 and 30% for validation subset, according to a modified multiple jack-knifing approach (Casa
301 et al., 2013a; Bishop & McBratney, 2001). The splitting was repeated 100 times, quantifying
302 the validation accuracy every time. The accuracy of the predictive models was finally
303 evaluated by examining the coefficient of determination (R^2), RMSE (Eq. (4)), Ratio of the
304 Performance to Deviation (RPD; Eq. (5)), and Ratio of Performance to Interquartile Range
305 (RPIQ; Eq. (6)). The equations used were as follows:

$$306 \quad RMSE = \sqrt{\frac{\sum_{i=1}^n (y_o - y_p)^2}{n}} \quad [4]$$

$$307 \quad RPD = \frac{sd}{RMSE} \quad [5]$$

$$308 \quad RPIQ = \frac{IQ}{RMSE} \quad [6]$$

309 where, y_o and y_p are the observed and predicted values, respectively, n is the number of data
310 pairs, and sd is the standard deviation of the observed values. We adopted threshold values of
311 RPD usually employed in the soil spectroscopy literature to assess the accuracy of soil
312 prediction models (Chang et al., 2001). These consider models with an excellent prediction
313 capability to have RPD values of >2 , intermediate capability to have values of $2-1.4$ and
314 unreliable model to have values of <1.4 . Recently, Bellon-Maurel et al. (2010) warned of the
315 reliability of these RPD thresholds because they were not statistically determined and the
316 standard deviation, used in the numerator of the RPD equation, does not describe the correct
317 range of variation for data with a skewed (non-normal) distribution. The RPIQ (using the IQ
318 instead of sd) provides a more robust statistic to describe the performance of estimation
319 models. For these reasons, both RPD and RPIQ values were reported to evaluate the
320 capability of the prediction models.

321 The one hundred RPIQ validation values of each PLSR model were used to detect statistical
322 differences among the estimation accuracies of the satellite imagers and between multispectral
323 and hyperspectral imagers for each soil variable. A one-way analysis of variance (ANOVA)
324 was applied to the RPIQ values of the one hundred replicates in order to verify that the
325 differences between RPIQ means were statistically significant ($P < 0.01$). Fisher's Least
326 Significant Difference (LSD) test was then applied to evaluate the statistical differences. The
327 results of the significance test for RPIQ can be considered equivalent to those that could be
328 obtained by testing differences in RMSE because of the relationship between the two statistics
329 (Eq.6).

330 The PLSR analyses were performed using the pls (Mevik et al., 2013) packages developed in
331 R software (R Development Core Team, 2011).

332 3. Results

333 3.1. Soil variables estimation using full spectra

334

335 The estimation of soil variables using the PONMAC full-resolution spectral library provided
336 good results: all soil texture variables were predicted with a RMSE lower than 5.4%, RPD
337 near to 2 and RPIQ of 2.51 for sand and higher than 3 for clay and silt (Table 5).

338 The statistical accuracy obtained using the LUCAS_C full-resolution spectral library was
339 lower compared with that obtained using the PONMAC library for all soil variables (Table 6).
340 In this case only the clay estimation model provided satisfactory prediction accuracy (Table 6;
341 RMSE = 7.63; RPD = 1.69; RPIQ = 2.34; $R^2 = 0.64$).

342 The VIP analysis highlighted the most important spectral regions for PLSR models of each
343 variable (Fig. 4). Clay, sand, silt and SOC estimation are highly affected by the wavelength
344 range between 1900 and 2400 nm ($VIP > 1$), but VIP values > 1 are also present in the VIS
345 region between 400 and 600 nm.

346

347 3.2. Soil variables estimation using resampled spectra

348

349 Compared with the full spectra PONMAC dataset, the accuracy of predictive models
350 generally decreased when using resampled spectra both for hyperspectral and multispectral
351 imagers, especially in the case of clay and silt estimation (Table 5). The RPIQ values obtained
352 from full spectrum data for clay (RPIQ = 3.07) and silt (RPIQ = 3.17) estimation were
353 significantly higher than those obtained with resampled spectra (Table 5), while no significant
354 differences were observed between full spectrum (RPIQ = 2.51) and all hyperspectral imagers

355 (RPIQ ranges from 2.45 and 2.56) for sand. The estimation accuracy of SOC for full spectrum
356 (RPIQ = 3.12), EnMAP (RPIQ = 3.11) and PRISMA (RPIQ = 3.02) data is statistically higher
357 as compared with the other imagers.

358 As expected, in the absence of noise, and taking into consideration only the spectral
359 characteristics (number, width and distribution of bands), the performances of the
360 forthcoming hyperspectral imagers are significantly better than those of multispectral imagers
361 and are very similar to Hyperion for all soil variables. An exception is observed for clay
362 estimation with Sentinel-2 data that showed results (RPIQ = 2.54) very similar to those
363 achieved using PRISMA and HypSIRI data (RPIQ = 2.43), while for sand, silt and SOC, the
364 differences with respect to hyperspectral imagers were statistically significant. Sentinel-2 data
365 showed the best performances among multispectral imagers for all variables, except for silt
366 (Table 5).

367 In the case of the LUCAS_C spectral library, differences between full-resolution and
368 resampled dataset were not always significantly in favor of hyperspectral sensors. The LSD
369 test highlighted an advantage of using simulated data instead of full spectrum data for the
370 estimation of soil texture components (Table 6). For example, the resampled hyperspectral
371 data (HypSIRI, EnMAP and PRISMA) showed significantly better results (RPIQ: 1.84–1.87)
372 compared with full spectrum data (RPIQ: 1.74) for silt estimation. Otherwise, a substantial
373 decrease of accuracy (compared with full spectra) is observed only for multispectral data
374 (Table 6) with the exception of silt estimation, for which the accuracy obtained by full
375 spectrum data was not statistically different from that obtained with ALI and Sentinel-2 data
376 (RPIQ: 1.71–1.76).

377 **3.3. Soil variables estimation using simulated spectra**

378

379 The use of simulated spectra, i.e. with the addition of noise and atmospheric effects, resulted
380 in a decrease of statistical accuracy in the estimation models, compared with full or resampled
381 spectra, both using PONMAC and LUCAS_C spectral libraries. Hyperion simulated data
382 appear particularly affected by noise, showing RPIQ values similar or lower than those
383 achieved using multispectral data (Table 7). Indeed, the LSD test carried out on the validation
384 statistics of the PONMAC dataset showed that the simulated Hyperion data provided similar
385 or worse results than multispectral imagers (Table 7). For clay estimation the best RPIQ
386 values were obtained using EnMAP, PRISMA and HypsIRI data (PONMAC: 2.16–2.35;
387 LUCAS_C: 1.95–2.00). The forthcoming hyperspectral imagers (EnMAP, PRISMA and
388 HypsIRI) display very similar capability levels when estimating soil texture and SOC content
389 both for the PONMAC (Table 7) and LUCAS_C (Table 8) dataset. Only clay content was
390 estimated with a suitable level of accuracy using PRISMA and HypsIRI resampled spectra of
391 LUCAS_C dataset (Table 8).

392 The difference in statistical accuracy between hyperspectral and multispectral data becomes
393 less consistent when using simulated data compared with resampled data. In many cases
394 multispectral data have similar or better results than hyperspectral, for example Landsat 8
395 provided an estimation accuracy of SOC estimation (RMSE = 0.25; RPD = 1.44; RPIQ =
396 2.48; $R^2 = 0.49$) not statistically different to hyperspectral imagers using PONMAC dataset.
397 Satisfactory estimation accuracy was also achieved by ALI when estimating clay (RPD =
398 1.45; RMSE = 6.58) and for all the multispectral imagers when estimating silt using
399 PONMAC dataset. However, for each soil variable, the LSD test confirmed better estimation
400 accuracy of the forthcoming hyperspectral imagers than for the current multispectral imagers.

401

402 **3.4. Sensitivity to bandwidth of the estimation models**

403

404 The issue of the low SNR of hyperspectral sensors due to the reduced energy collected by the
405 sensors in narrow spectral bands (evidenced by the results of the previous section) was
406 investigated by examining the effect of broadening the bandwidths. The application of PLSR
407 models to the PONMAC datasets showed, for all variables only a slight decrease of
408 estimation accuracy when changing bandwidth from 10 to 40 nm (Fig. 5) in the case of
409 absence of noise (resampled dataset). A greater decrease of the RPD values is observed when
410 using a bandwidth greater than 80 nm. After the introduction of noise (simulated dataset), the
411 SOC estimation accuracy is almost insensitive to the reduction of spectral resolution, while
412 RPD values of clay and sand even increase slightly up to 40 nm (Fig. 6). Furthermore, the
413 RPD values only decrease significantly when using a bandwidth greater than 80 nm.

414

415 **4. Discussion**

416

417 The statistical accuracy obtained using the LUCAS_C spectral library was generally lower
418 than that obtained with the PONMAC dataset for full spectrum, resampled and simulated
419 data. These differences in model accuracy are probably due to the high variability of the
420 LUCAS_C dataset (Table 1), and to the different methodologies adopted for soil analyses and
421 spectra acquisition. Although soil samples of the PONMAC dataset cover a wide range of
422 clay, sand, silt and SOC values they are characterized by two types of parental material. On
423 the other hand, LUCAS_C represents most of the agricultural soils of Italy. This high
424 variability includes different compositions of clay minerals and organic matter among soil
425 samples, which influences the number and diversity of PLSR components. In this regard, the
426 typical absorption features related to the clay lattice are different according to the
427 predominant clay mineral: the absorption features of kaolinite and montmorillonite are
428 located at 1400, 1900, and 2200 nm (Clark, 1999), whereas illite has two bands around 2300

429 and 2400 nm. The typical absorption features of organic matter are distributed throughout the
430 entire reflectance VNIR–SWIR spectrum. This is due to its heterogeneous composition, in
431 turn derived from varied source materials and their mineralization levels (Ben-Dor et al.,
432 1997; Li et al., 2012; Mouazen Maleki et al., 2007). Therefore, the heterogeneous
433 composition of the soil samples of the LUCAS_C dataset involves the use of more PLSR
434 components than those used for the PONMAC dataset (Table 5; Table 6).

435 The VIP analysis confirmed the most important spectral region ($VIP > 1$) for clay, sand, silt
436 and SOC is located between 1900 and 2400 nm. This result can be attributed mainly to the
437 presence of typical absorption features in the SWIR spectral region. In particular, metal-OH
438 bending and stretching of O-H bonds related to the clay lattice affect the absorption peaks
439 between 2160 and 2300 nm (Ben-Dor et al., 2009); while VIP scores of SOC in the SWIR are
440 mainly affected by phenolic, amide and aliphatic groups between 2200 and 2220 nm. The
441 high VIP values around 1900 nm are due to the absorption features of water. Viscarra-Rossel
442 and Beherens (2010) used VIP values to select the most useful bands for estimating clay and
443 SOC by means of diffuse reflectance spectra. Most of the selected bands they identified, for
444 both variables, were located between 1900 and 2400 nm, similarly to our results.

445 VIP values greater than 1 are present also in the VIS region between 400 and 600 nm.
446 Viscarra-Rossel et al. (2008) detected high VIP values between 400 and 600 nm for SOC
447 estimation, using spectra acquired by an infrared spectrometer from 400 to 1000 nm.
448 Although some wavelengths correlated with organic matter exist at 450 nm and 590 nm
449 (Melendez-Pastor et al., 2008; Li et al., 2012), the main information provided by the VIS
450 region, for texture and SOC estimation, is likely to be related to the hue of soil. Since soils
451 having dark hues, for equal moisture content and parental material, generally contain higher
452 clay and SOC and lower sand content than soils with pale hues. The consistent worsening of
453 the estimation accuracy when using simulated hyperspectral data, instead of resampled

454 spectra, highlights the importance of the effect of the noise affecting the various instruments.
455 In general the noise particularly affects the SWIR region, especially from 2000 nm to 2400
456 nm (Fig. 1 and Fig. 2), where the VIP analysis provided the higher values (Fig. 4). In
457 addition, the very low SNR that corresponds to water absorption features required the
458 elimination of the corresponding spectral bands.

459 The estimation accuracy showed that the addition of noise to the spectra reduces the
460 differences of performance between hyperspectral and multispectral imagers. However the
461 LSD test detected a significant difference between hyperspectral (with exclusion of Hyperion)
462 and multispectral imagers, for both spectral libraries and each soil variable. Furthermore, the
463 accuracy of soil variable estimation is very similar for the three forthcoming imagers (Table 7
464 and Table 8), in spite of the fact that EnMAP and PRISMA present (in sensitive spectral
465 ranges) a greater number of bands (but a lower SNR) than HypsIRI (Table 9). These results
466 are explainable if the SNR, which generally decreases with the spectral resolution due to the
467 lower energy reaching the sensors, partially compensates the benefits of the higher spectral
468 resolution, thus better exploiting narrow spectral features. Obviously, spatial resolution
469 influences, *ceteris paribus*, the prediction accuracy of soil variables (Gomez et al., 2015). For
470 this reason, the estimation accuracy obtained by HypsIRI would not be perfectly comparable
471 with EnMap and PRISMA data, because the US imager will have a spatial resolution much
472 coarser (GSD of 60 m) than the other hyperspectral sensors investigated in this work (GSD of
473 30 m).

474 The analysis concerning the sensitivity to bandwidth of the estimation accuracy of soil
475 variables showed that data acquired between 400 and 2400 nm, with a bandwidth of 40 nm
476 could provide results very similar to those obtained using sensors with a higher spectral
477 resolution, because of the lower SNR of the former. However, a bandwidth of 40 nm could be
478 useful for multivariate prediction models, such as the PLSR, where the spectral information is

479 summarized in a restricted number of regressors (PLSR components) correlated with the
480 target variable. Actually, PLSR technique has proved less effective than other hybrid
481 estimation techniques, such as regression kriging (Casa et al., 2013a) or linear mixed effect
482 models (Castaldi et al., 2014); however these hybrid techniques cannot be used with
483 laboratory samples because they take into account the spatial correlation between soil
484 samples.

485 The higher spectral resolution of hyperspectral data permits exploitation of well-defined
486 narrow absorption spectral features typically associated to overtones of functional groups in
487 the VNIR and SWIR spectral ranges related to the estimation of some chromophores (Ben-
488 Dor et al., 2009), such as the clay minerals (Clark, 1999). Unfortunately, the sensor's noise
489 generally increases with increasing spectral resolution. Thus, as confirmed by the analysis of
490 the sensitivity to bandwidth, the advantages of detecting specific narrow bands could be
491 nullified by low SNR, especially in the SWIR spectral region where the absorption peaks of
492 the main soil variables are located (Lobell & Asner, 2002).

493 This work provides an evaluation of the capability of the remote imagers to estimate soil
494 variables using simulated spectra, taking into account the spectral resolution and noise of each
495 imager and the atmospheric effect. Both PONMAC and LUCAS_C spectral libraries are
496 composed by spectra acquired in laboratory conditions on air dried soil samples, having an
497 artificial roughness; however the actual remote sensing data are also affected by the
498 confounding effects of soil moisture, soil roughness and crop residues. Soil moisture has a
499 strong influence on the amount and composition of reflected and emitted energy from a soil
500 surface, reducing the reflectance over the entire spectrum (Lobell & Asner, 2002; Nocita et
501 al., 2013; Rienzi et al. 2014; Castaldi et al. 2015), and the soil roughness influences the
502 scattering effect. The drying process of soil samples of both datasets used in the present work
503 has occurred naturally and for this reason soil datasets preserved moisture variability which

504 could be assumed to be similar to that observable on the top layer of an agricultural soil, with
505 quite a homogenous roughness, after a period of dry weather. In this regard there are many
506 examples in the literature of the estimation of soil variables using remote data which assume
507 the homogeneity of soil roughness of the bare soil after seed bed preparation, thus without
508 crop residues or weeds within the field (Casa et al., 2013b; Ge et al., 2011; Mulder et al.,
509 2011; Selige et al., 2006). Nevertheless, the capability of the present and forthcoming satellite
510 imagers shall be assessed using actual data. Effective comparisons can only be done using
511 satellite data acquired for a particular field site over a particular time period. In this regard,
512 Castaldi et al. (2014) compared Hyperion and ALI imager using images acquired on the same
513 date and obtained more accurate prediction with Hyperion as compared with ALI for clay,
514 sand and soil organic matter. However, at present, the comparison among current and
515 forthcoming sensors is only possible using spectral laboratory data covering the whole VIS–
516 SWIR spectral range (e.g., PONMAC and LUCAS_C), considering that a satellite
517 hyperspectral sensor having spectral characteristics suitable to simulate PRISMA, EnMAP or
518 HypsIRI images does not exist yet.

519

520 **5. Conclusions**

521

522 This work investigates, for seven satellite imagers, the effects of spectral resolution and SNR
523 on the estimation of some soil variables such as clay, sand, silt and SOC by using PLSR
524 models. We compared the estimation accuracy of soil texture and SOC using resampled and
525 simulated spectra according to spectral and noise characteristics of four current (ALI, Landsat
526 8, Sentinel-2 and Hyperion) and three forthcoming (EnMAP, PRISMA, HypsIRI) satellite
527 imagers. The LSD testing on simulated data (Table 8) suggests that the forthcoming
528 hyperspectral imagers will enhance, in relative terms, the accuracy of soil variable estimation

529 from bare soil imagery, as compared to current generation sensors, though in absolute terms
530 the statistics of their performance (e.g. in terms of r^2) was not always satisfactory. The
531 advancement of the new generation hyperspectral imagers, as compared to Hyperion, is due to
532 the improvement of the SNR in the SWIR region, in particular between 2000 and 2400 nm.
533 EnMAP, PRISMA and HypsIRI imagers provided significantly better estimation accuracy
534 than ALI, Landsat 8 and Sentinel-2. This is mainly due to the higher number of bands,
535 especially in SWIR region, and the narrower bandwidths that allow them to better exploit the
536 spectral features of clay minerals and organic matter. However, the benefit of improved
537 spectral resolution is partially offset by the amplification of noise when increasing the number
538 of spectral bands. Analysis of the sensitivity to bandwidth highlighted that a bandwidth of 40
539 nm could provide PLSR results very similar to those obtained with sensors having a higher
540 spectral resolution and that, only when the noise levels are decreased, the capabilities of
541 current and forthcoming hyperspectral sensors will be fully achieved. The analysis also
542 indicates that 40 nm is a maximum value for the bandwidth, beyond which the estimation
543 accuracy of soil texture sharply decreases.

544 We conclude that hyperspectral data from the forthcoming satellite missions could be more
545 valuable for mapping and monitoring soil texture and SOC as compared to current imagers.
546 The higher spectral resolution of the new imagers, coupled to an improved SNR and an
547 extensive spatial coverage and frequent revisit period, will guarantee highly detailed spectral
548 information, suitable to characterize and monitor surface soil variables and soil erosion
549 processes. Nevertheless, as shown in this work, the achievement of satisfactory quantitative
550 estimation results from hyperspectral imagers is still hampered by a low SNR in the SWIR
551 spectral region, especially apparent when using a large and varied dataset (LUCAS_C). On
552 the other hand, the use of hyperspectral data joined with spatial predictive models on
553 delimited field data (PONMAC) could lead to a satisfactory estimation of soil variables. [rc1]

554 Further research will need to focus on estimation accuracies for the retrieval of soil variables
555 using, once available, real data from next generation hyperspectral satellite imagers.

556

557

558 **Acknowledgements**

559 The research activities have been co-funded by the Italian Space Agency (ASI) with grant
560 I/019/11/10 (PRISMA Mission). Additionally, we would like to thank to the PRISMA,
561 EnMAP and HypIRI teams for having provided information on the radiometric
562 characteristics of the sensors useful for the production of the datasets used in this paper.
563 Additionally we are grateful to the anonymous reviewers for their valuable comments, which
564 helped to improve the quality of this paper.

565

566

567 **References**

568 Baillarin, S.J., Meyret, A., Dechoz, C., Lacherade, S., Tremas, T., Isola, C., ... Spoto, F.
569 (2012). Sentinel-2 level 1 products and image processing performances. Proceedings of
570 Geoscience and Remote Sensing Symposium (IGARSS), 2012 IEEE International ,
571 pp.7003–006, 22–27 July 2012.

572 Bellon-Maurel, V., Fernandez-Ahumada, E., Palagos, B., Roger, J.M., & McBratney, A.
573 (2010). Critical review of chemometric indicators commonly used for assessing the
574 quality of the prediction of soil attributes by NIR spectroscopy. *TrAC Trends in*
575 *Analytical Chemistry*, 29(9), 1073–1081. doi:10.1016/j.trac.2010.05.006.

576 Ben-Dor, E., Inbar, Y., Chen, Y. (1997). The reflectance spectra of organic matter in the
577 visible near-infrared and short wave infrared region (400–2500 nm) during a controlled
578 decomposition process. *Remote Sensing of Environment*, 61(1), 1–15.
579 [http://doi.org/10.1016/S0034-4257\(96\)00120-4](http://doi.org/10.1016/S0034-4257(96)00120-4)

580 Ben-Dor, E., Chabrilat, S., Demattê, J.A.M., Taylor, G.R., Hill, J., Whiting, M.L., &
581 Sommer, S. (2009). Using Imaging Spectroscopy to study soil properties. *Remote*
582 *Sensing of Environment*, 113, S38–S55. <http://doi.org/10.1016/j.rse.2008.09.019>

583 Ben-Dor, E., Kafri A., & Varacalli G. (2013). SHALOM: Spaceborne Hyperspectral
584 Applicative Land and Ocean Mission: A joint project of ASI – ISA. Proceedings of the
585 International Geoscience and Remote Sensing Symposium (IAGARSS’13), Melbourne,
586 Australia, 4 pages, 2013.

587 Bishop, T.F., & McBratney, A. (2001). A comparison of prediction methods for the creation
588 of field-extent soil property maps. *Geoderma*, 103(1–2), 149–160. doi:10.1016/S0016-
589 7061(01)00074-X

590 Casa, R., Castaldi, F., Pascucci, S., Basso, B., & Pignatti, S. (2013a). Geophysical and
591 hyperspectral data fusion techniques for in-field estimation of soil properties. *Vadose*
592 *Zone Journal*, 12(4). doi:10.2136/vzj2012.0201

593 Casa, R., Castaldi, F., Pascucci, S., Palombo, A., & Pignatti, S. (2013b). A comparison of
594 sensor resolution and calibration strategies for soil texture estimation from hyperspectral
595 remote sensing. *Geoderma*, 197–198, 17–26. doi:10.1016/j.geoderma.2012.12.016

596 Castaldi, F., Casa, R., Castrignanò, A., Pascucci, S., Palombo, A., & Pignatti, S. (2014).
597 Estimation of soil properties at the field scale from satellite data: a comparison between

598 spatial and non-spatial techniques. *European Journal of Soil Science*, 65(6), 842–851.
599 doi:10.1111/ejss.12202

600 Castaldi, F., Palombo, A., Pascucci, S., Pignatti, S., Santini, F., & Casa, R. (2015). Reducing
601 the Influence of Soil Moisture on the Estimation of Clay from Hyperspectral Data: A
602 Case Study Using Simulated PRISMA Data. *Remote Sensing*, 7(11), 15561–15582.
603 doi:10.3390/rs71115561

604 Chang C. I. (2007). *Hyperspectral Data Exploitation: Theory and Applications*. Hoboken,
605 NJ: Wiley.

606 Chang, C.W., Laird, D.A., Mausbach, M.J., & Hurburgh, C.R. (2001). Near-Infrared
607 Reflectance Spectroscopy – Principal components regression analyses of soil properties.
608 *Soil Science Society of America Journal*, 65(2), 480. doi:10.2136/sssaj2001.652480x

609 Chen, L., Flynn, D.F.B., Jing, X., Kühn, P., Scholten, T., & He, J.S. (2015). A comparison of
610 two methods for quantifying soil organic carbon of alpine grasslands on the Tibetan
611 Plateau. *PLOS ONE*, 10(5), e0126372. doi:10.1371/journal.pone.0126372

612 Chong, I.G., & Jun, C.H. (2005). Performance of some variable selection methods when
613 multicollinearity is present. *Chemometrics and Intelligent Laboratory Systems*, 78(1-2),
614 103–112

615 Clark, R.N. 1999. Chapter 1: Spectroscopy of Rocks and Minerals, and Principles of
616 Spectroscopy, in: Rencz A.N. (Eds.), *Manual of Remote Sensing*, John Wiley and Sons,
617 New York, pp 3- 58.

618 Costantini, E.A.C., L'Abate, G., Barbetti, R., Fantappiè, M., Lorenzetti, R. and Magini, S.
619 (2012). Carta dei suoli d'Italia, scala 1:1.000.000 (Soil map of Italy, scale 1:1.000.000).

620 Consiglio per ricerca e la sperimentazione in agricoltura, S.EL.CA. Florence/Firenze,
621 Italy (ISBN: 978-88-97002-02-4). Available at: <http://www.soilmaps.it/>

622 Demattê, J.A.M., Fiorio, P.R., & Ben-Dor, E. (2009). Estimation of soil properties by orbital
623 and laboratory reflectance means and its relation with soil classification. *The Open*
624 *Remote Sensing Journal*, 2(1), 12–23. doi:10.2174/1875413900902010012

625 FAO-ISRIC-ISSS. (1998). World Reference Base for Soil Resources. World Soil Resources
626 Report 84. Food and Agriculture Organisation, Rome, Italy.

627 Ge, Y., Thomasson, J. A., & Sui, R. (2011). Remote sensing of soil properties in precision
628 agriculture: A review. *Frontiers of Earth Science*, 5(3), 229 - 238. doi:10.1007/s11707-
629 011-0175-0

630 Gomez, C., Oltra-Carrió, R., Bacha, S., Lagacherie, P., & Briottet, X. (2015). Evaluating the
631 sensitivity of clay content prediction to atmospheric effects and degradation of image
632 spatial resolution using Hyperspectral VNIR/SWIR imagery. *Remote Sensing of*
633 *Environment*, 164, 1–15. <http://doi.org/10.1016/j.rse.2015.02.019>

634 Gomez, C., Viscarra Rossel, R.A., & McBratney, A.B. (2008). Soil organic carbon prediction
635 by hyperspectral remote sensing and field vis-NIR spectroscopy: An Australian case
636 study. *Geoderma*, 146(3-4), 403–411. doi:10.1016/j.geoderma.2008.06.011

637 Green, R.O., Asner, G., Ungar, S., & Knox, R. (2008). NASA mission to measure global
638 plant physiology and functional types. Proceedings in Aerospace Conference, 2008
639 IEEE, pp.1–7, 1–8 March 2008. doi: 10.1109/AERO.2008.4526244

640 Houborg, R., Anderson, M., Gao, F., Schull, M., & Cammalleri, C. (2012). Monitoring water
641 and carbon fluxes at fine spatial scales using HypSPIRI-like measurements. In *2012 IEEE*

642 *International Geoscience and Remote Sensing Symposium*, pp. 7302–7305. IEEE.
643 doi:10.1109/IGARSS.2012.6351975

644 International Standards Organisation 10694 (1995). Soil quality — Determination of organic
645 and total carbon after dry combustion (elementary analysis)

646 International Standards Organisation 11277 (1998). Soil quality — Determination of particle
647 size distribution in mineral soil material — Method by sieving and sedimentation.

648 Kaufmann, Y.J., Wald, A.E., Remer, L.A., Gao, B.C., Li, R.R., & Flynn, L. (1997). The
649 MODIS 2.1-mm channel-correlation with visible reflectance for use in remote sensing of
650 aerosol. *Proceedings in IEEE Transactions on Geoscience and Remote Sensing*. Vol. 35,
651 no. 5, pp. 1286–1298.

652 Lagacherie, P., Baret, F., Feret, J.B., Netto, J.M., & Robbez-Masson, J.M. (2008). Estimation
653 of soil clay and calcium carbonate using laboratory, field and airborne hyperspectral
654 measurements, *Remote Sensing of Environment*, 112, 825–835.
655 doi:10.1016/j.rse.2007.06.014.

656 Li, D., Chen, X., Peng, Z., Chen, S., Chen, W., Han, L., & Li, Y. (2012). Prediction of soil
657 organic matter content in a litchi orchard of South China using spectral indices. *Soil and*
658 *Tillage Research*, 123, 78–86. doi:10.1016/j.still.2012.03.013

659 Lobell, D.B., & Asner, G.P. (2002). Moisture effects on soil reflectance. *Soil Science Society*
660 *of America Journal*, 66(3), 722–727. Retrieved from
661 [http://www.scopus.com/inward/record.url?eid=2-s2.0-](http://www.scopus.com/inward/record.url?eid=2-s2.0-0036240171&partnerID=tZOtx3y1)
662 [0036240171&partnerID=tZOtx3y1](http://www.scopus.com/inward/record.url?eid=2-s2.0-0036240171&partnerID=tZOtx3y1)

663 Mulder, V.L., de Bruin, S., Schaepman, M.E., & Mayr, T.R. (2011). The use of remote
664 sensing in soil and terrain mapping — A review. *Geoderma*, 162(1-2), 1–19.
665 doi:10.1016/j.geoderma.2010.12.018

666 Markham, B.L., Knight, E.J., Canova, B., Donley, E., Kvaran, G., Lee, K., Barsi, J.A.,
667 Pedelty, J.A., Dabney, P.W., & Irons, J.R. (2012). The Landsat data continuity mission
668 Operational Land Imager (OLI) sensor. Proceedings in Geoscience and Remote Sensing
669 Symposium (IGARSS), 2012 IEEE International, pp.6995–6998, 22–27 July 2012. doi:
670 10.1109/IGARSS.2012.6351961

671 Melendez-Pastor, I., Navarro-Pedreño, J., Gómez, I., & Koch, M. (2008). Identifying optimal
672 spectral bands to assess soil properties with VNIR radiometry in semi-arid soils.
673 *Geoderma*, 147(3), 126–132.

674 Mendenhall, J.A., Hearn, D.R., Evans, J.B, Lencioni, D.E., Digenis, C.J., & Welsh, R.D.
675 (2001) Initial flight test results from the EO-1 Advanced Land Imager: radiometric
676 performance. Proceedings in Geoscience and Remote Sensing Symposium, 2001.
677 IGARSS '01. IEEE 2001 International , vol.1, pp.515 -517, IEEE. doi:
678 10.1109/IGARSS.2001.976207

679 Mevik, B.H. & Wehrens, R. 2007. The pls package: Principal component and partial least
680 squares regression in R. *Journal of Statistical Software*, 18, 1–23.

681 Michel, S., Gamet, P., & Lefevre-Fonollosa, M.J. (2011). HYPXIM — A hyperspectral
682 satellite defined for science, security and defence users. IEEE Proceedings in 2011 3rd
683 *Workshop on Hyperspectral Image and Signal Processing: Evolution in Remote Sensing*
684 (*WHISPERS*) (pp. 1–4). doi:10.1109/WHISPERS.2011.6080864

685 Mouazen, A.M., Maleki, M.R., De Baerdemaeker, J., & Ramon, H. (2007). On-line
686 measurement of some selected soil properties using a VIS–NIR sensor. *Soil and Tillage*
687 *Research*, 93(1), 13–27. doi:10.1016/j.still.2006.03.009.

688 Mulder, V.L., de Bruin, S., Schaepman, M.E., & Mayr, T. R. (2011). The use of remote
689 sensing in soil and terrain mapping — A review. *Geoderma*, 162(1-2), 1–19.
690 doi:10.1016/j.geoderma.2010.12.018

691 Nocita, M., Stevens, A., Noon, C., & van Wesemael, B. (2013). Prediction of soil organic
692 carbon for different levels of soil moisture using Vis-NIR spectroscopy. *Geoderma*, 199,
693 37–42. <http://doi.org/10.1016/j.geoderma.2012.07.020>

694 Odeh, I.O., & McBratney, A.B. (2000). Using AVHRR images for spatial prediction of clay
695 content in the lower Namoi Valley of eastern Australia. *Geoderma*, 97(3–4), 237–254.
696 doi:10.1016/S0016-7061(00)00041-0

697 Pascucci, S., Casa, R., Belviso, C., Palombo, A., Pignatti, S., & Castaldi, F. (2014).
698 Estimation of soil organic carbon from airborne hyperspectral thermal infrared data: a
699 case study. *European Journal of Soil Science*, 65(6), 865–875. doi:10.1111/ejss.12203

700 Pignatti, S., Acito, N., Amato, U., Casa, R., de Bonis, R., Diani, M., ... Cuomo, V. (2012).
701 Development of algorithms and products for supporting the Italian hyperspectral
702 PRISMA mission: The SAP4PRISMA project. Proceedings in *2012 IEEE International*
703 *Geoscience and Remote Sensing Symposium* (pp. 127–130). IEEE.
704 doi:10.1109/IGARSS.2012.6351620

705 Pignatti, S., Palombo, A., Pascucci, S., Romano F., Santini F., Simoniello T., ... Ananasso C.
706 (2013). The PRISMA hyperspectral mission: Science activities and opportunities for
707 agriculture and land monitoring. Proceedings in *Geoscience and Remote Sensing*

708 *Symposium (IGARSS), 2013 IEEE International* , pp.4558 -4561, 21-26 July 2013. doi:
709 10.1109/IGARSS.2013.6723850.

710 R Development Core Team 2011. R: A Language and Environment for Statistical Computing.
711 R Foundation for Statistical Computing, Vienna. ISBN 3-900051-07-0 [WWW
712 document]. Retrieved from: <http://www.R-project.org/>.

713 Richter, K., Hank, T., Atzberger, C., Locherer, M., & Mauser, W. (2012). Regularization
714 strategies for agricultural monitoring: The EnMAP vegetation analyzer (AVA).
715 Proceedings in *2012 IEEE International Geoscience and Remote Sensing Symposium*,
716 pp. 6613–6616. IEEE. doi:10.1109/IGARSS.2012.6352083

717 Rienzi, E.A., Mijatovic, B., Mueller, T.G., Matocha T.J., Sikora F.J., & Castrignanò A.
718 (2014). Prediction of soil organic carbon under varying moisture levels using reflectance
719 spectroscopy. *Soil Science Society of America Journal*, 78(3), 958.
720 <http://doi.org/10.2136/sssaj2013.09.0408>

721 Sang, B., Schubert, J., Kaiser, S., Mogulsky V., Neumann C., Förster K.P., ... Chlebek C.
722 (2008). The EnMAP hyperspectral imaging spectrometer:instrument concept, calibration
723 and technologies. Proceedings in SPIE 7086, Imaging Spectrometry XIII, 708605 (27
724 August 2008). doi: 10.1117/12.794870

725 Selige, T., Böhner, J., & Schmidhalter, U. (2006). High resolution topsoil mapping using
726 hyperspectral image and field data in multivariate regression modeling procedures.
727 *Geoderma*, 136(1–2), 235–244. doi:10.1016/j.geoderma.2006.03.050

728 Soil Survey Staff 2014. Keys to Soil Taxonomy, 12th edn. USDA-Natural Resources
729 Conservation Service, Washington.

730 SPS (2011). The Soil Spectroscopy Group. Available at: <http://groups.google.com/group/soil->
731 [spectroscopy](http://groups.google.com/group/soil-spectroscopy)

732 Stevens, A., Nocita, M., Tóth, G., Montanarella, L., & van Wesemael, B. (2013). Prediction
733 of soil organic carbon at the European scale by visible and near infrared reflectance
734 spectroscopy, *PLoS ONE*, 8(6), 1–13.

735 Tanii, J., Iwasaki, A., Kawashima, T., & Inada, H. (2012). Results of evaluation model of
736 Hyperspectral Imager Suite (HISUI). Proceedings in *2012 IEEE International*
737 *Geoscience and Remote Sensing Symposium*, pp. 131–134. IEEE.
738 doi:10.1109/IGARSS.2012.6351619

739 Tóth, G., Jones, A., & Montanarella, L. (2013). The LUCAS topsoil database and derived
740 information on the regional variability of cropland topsoil properties in the European
741 Union. *Environmental Monitoring and Assessment*, 185(9), 7409–25.
742 doi:10.1007/s10661-013-3109-3

743 Tukey, J. W. (1977). Some thoughts on clinical trials, especially problems of multiplicity.
744 *Science*, 198(4318), 679–684. Retrieved from
745 [http://www.scopus.com/inward/record.url?eid=2-s2.0-](http://www.scopus.com/inward/record.url?eid=2-s2.0-0017778629&partnerID=tZOtx3y1)
746 [0017778629&partnerID=tZOtx3y1](http://www.scopus.com/inward/record.url?eid=2-s2.0-0017778629&partnerID=tZOtx3y1)

747 Vermote, E.F., Tanre, D., Deuze, J.L., Herman, M., & Morcette, J.J. (1997). Second
748 simulation of the satellite signal in the solar spectrum, 6S: an overview. *IEEE*
749 *Transactions on Geoscience and Remote Sensing*, 35(3), 675–686.
750 doi:10.1109/36.581987

- 751 Viscarra Rossel, R. A., & Behrens, T. (2010). Using data mining to model and interpret soil
752 diffuse reflectance spectra. *Geoderma*, 158(1-2), 46–54.
753 doi:10.1016/j.geoderma.2009.12.025
- 754 Viscarra Rossel, R. A., Fouad, Y., & Walter, C. (2008). Using a digital camera to measure
755 soil organic carbon and iron contents. *Biosystems Engineering*, 100(2), 149–159.
756 doi:10.1016/j.biosystemseng.2008.02.007
- 757 Willoughby, C.T., Marmo, J., & Folkman, M.A. (1996). Hyperspectral imaging payload for
758 the NASA small satellite technology initiative program, Proceedings in Aerospace
759 Applications Conference, vol.2, pp.67–79. IEEE. doi: 10.1109/AERO.1996.495966
- 760 Wold, S., Sjöström, M., & Eriksson, L. (2001). PLS-regression: a basic tool of chemometrics.
761 *Chemometrics and Intelligent Laboratory Systems*, 58(2), 109–130. Retrieved from
762 <http://www.sciencedirect.com/science/article/pii/S0169743901001551>
- 763 Wu, J., Li, Z., Gao, Z., Wang, B., Bai, L., Sun, B., ... Ding, X. (2015). Degraded land
764 detection by soil particle composition derived from multispectral remote sensing data in
765 the Otindag Sandy Lands of China. *Geoderma*, 241-242, 97–106.
766 doi:10.1016/j.geoderma.2014.11.011
- 767 Zhai, Y., Thomasson, J.A., Boggess, J.E., & Sui, R. (2006). Soil texture classification with
768 artificial neural networks operating on remote sensing data. *Computers and Electronics
769 in Agriculture*, 54(2), 53–68. doi:10.1016/j.compag.2006.08.001
- 770 Zhang, T., Li, L., & Zheng, B. (2013). Estimation of agricultural soil properties with imaging
771 and laboratory spectroscopy. *Journal of Applied Remote Sensing*, 7(1), 073587.
772 doi:10.1117/1.JRS.7.073587

774 **Figure captions**

775

776 Figure 1. HypsIRI spectral noise characteristics provided by HypsIRI scientific team. The
777 SNR spectra (left) are shown for four different radiance levels (right), corresponding to
778 different albedo and solar zenith angles.

779 Figure 2. Hyperion simulated spectrum obtained from the PONMAC library without noise
780 (dashed line) and with noise (solid line).

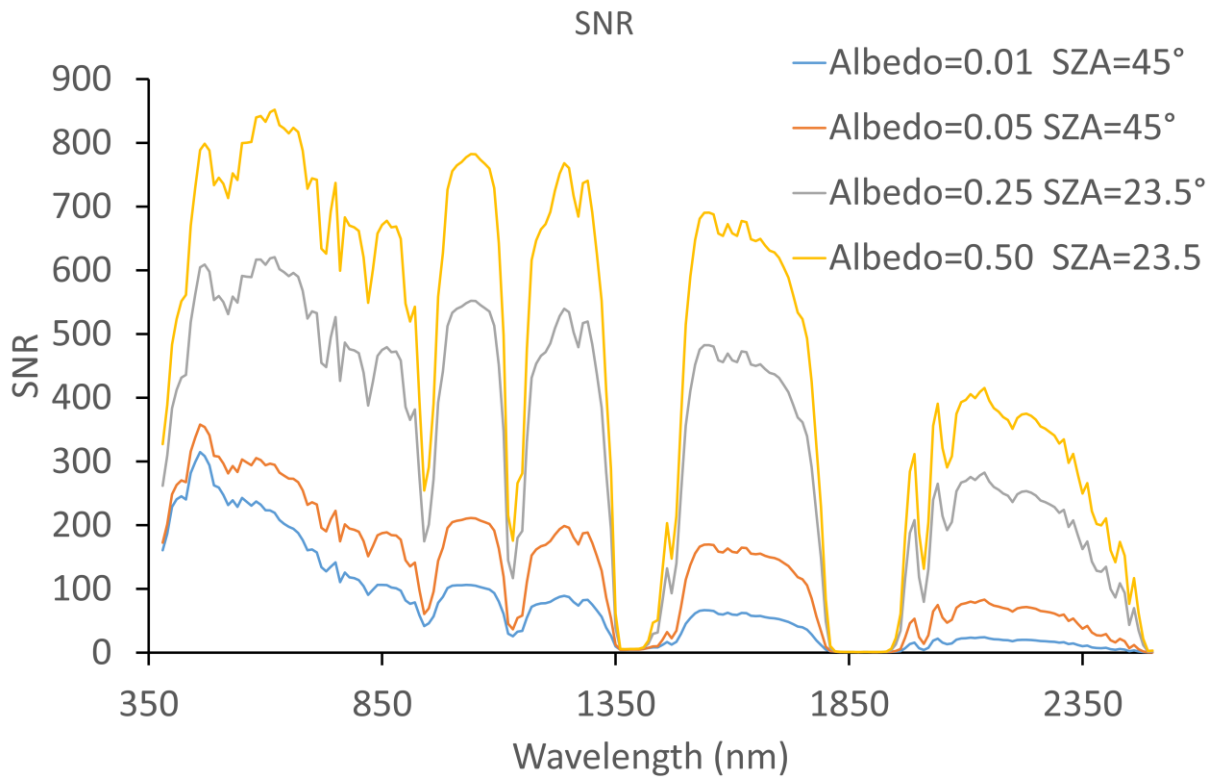
781 Figure 3. Set of five spectra obtained by resampling a simulated HypsIRI soil spectrum,
782 derived from the PONMAC dataset, to bandwidths from 10 nm to 160 nm. The spectra were
783 progressively offset with a factor of 0.1 for visualization purposes. Noisy bands in the water
784 absorption ranges were removed.

785 Figure 4. Values of Variable Importance in Projection (VIP) in PLSR estimation models for
786 clay (a–b), sand (c–d), silt (e–f) and SOC (g–h) using full spectrum data of PONMAC (left
787 column) and LUCAS_C (right column) library. Predictors with VIP values greater than one
788 (red area above the horizontal line in the plots) were considered significant for the PLSR
789 models.

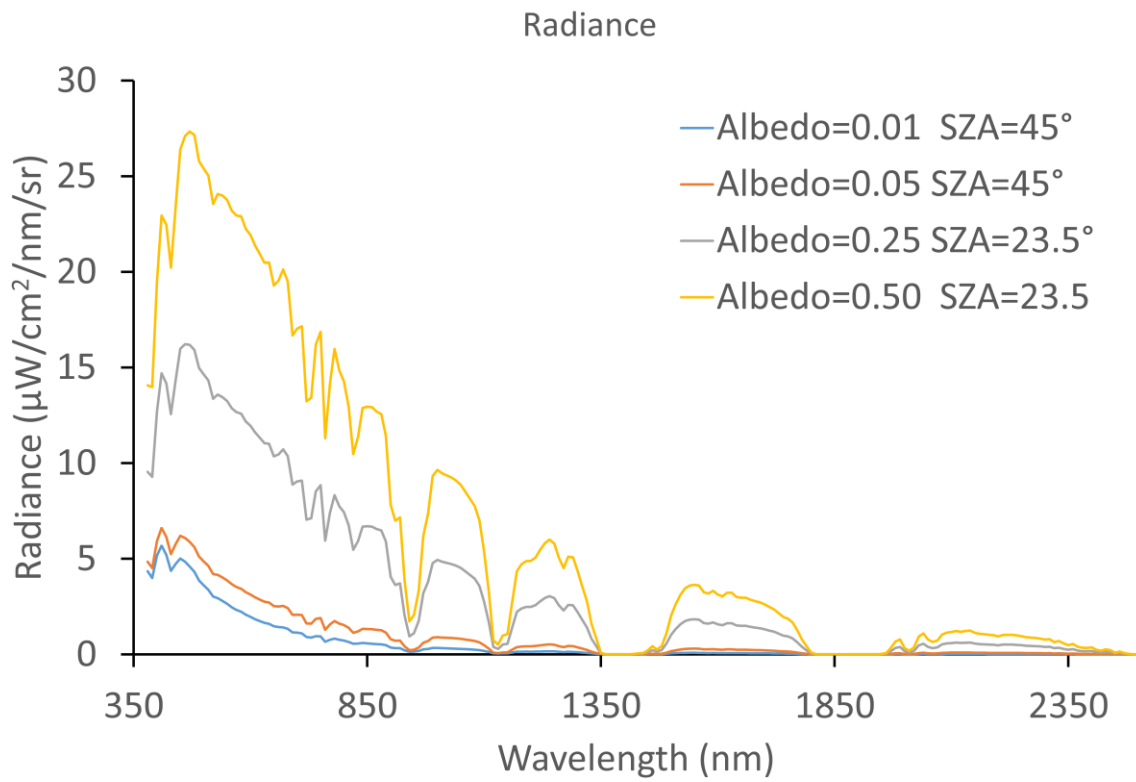
790 Figure 5. Ratio of performance to deviation (RPD) of clay, sand, silt and SOC estimation,
791 obtained from the degradation of the spectral resolution of *resampled* spectra from 10 to 20,
792 40, 80 and 160 nm bandwidth.

793 Figure 6. Ratio of performance to deviation (RPD) of clay, sand, silt and SOC estimation,
794 obtained from the degradation of the spectral resolution of *simulated* spectra from 10 to 20,
795 40, 80 and 160 nm bandwidth.

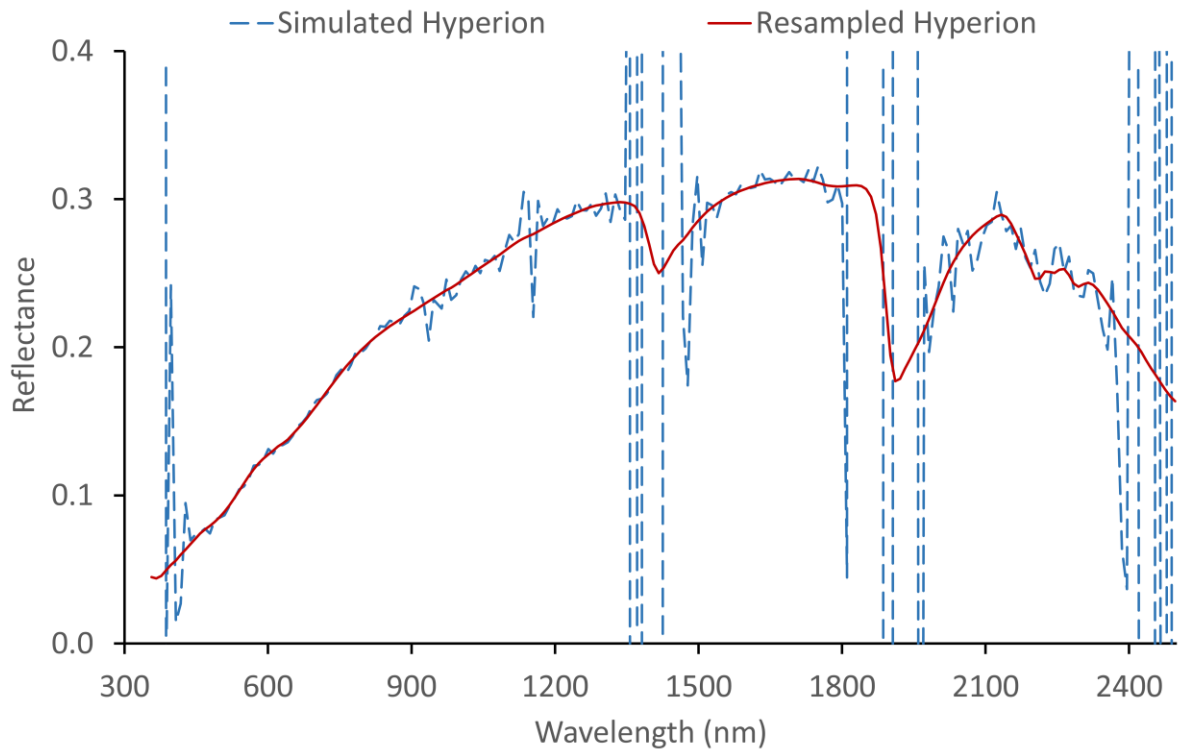
796



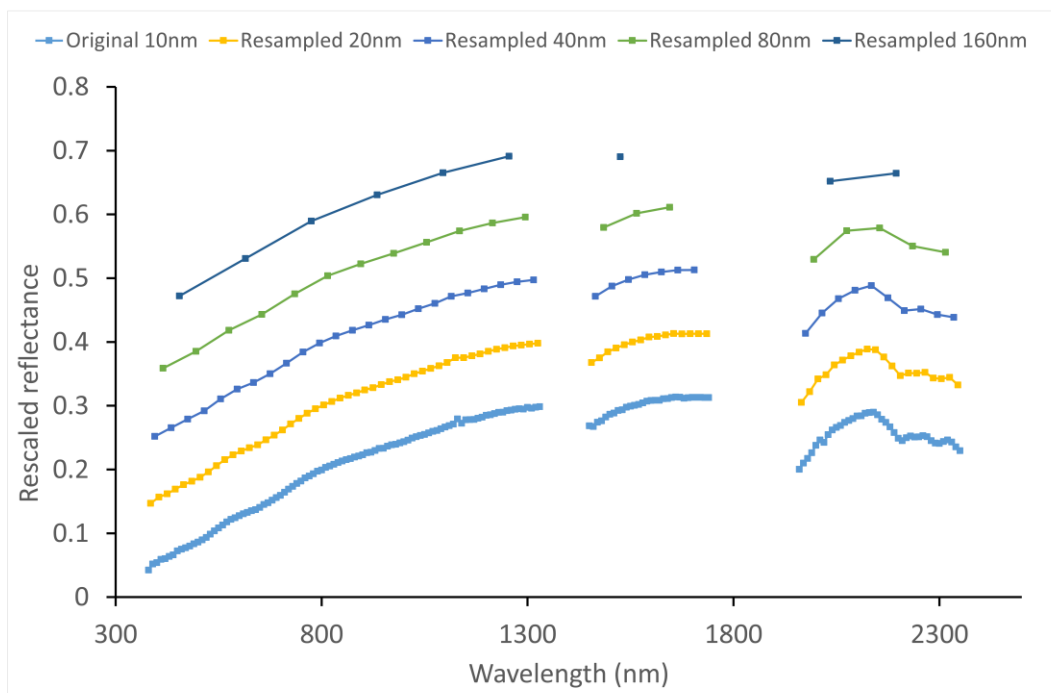
797



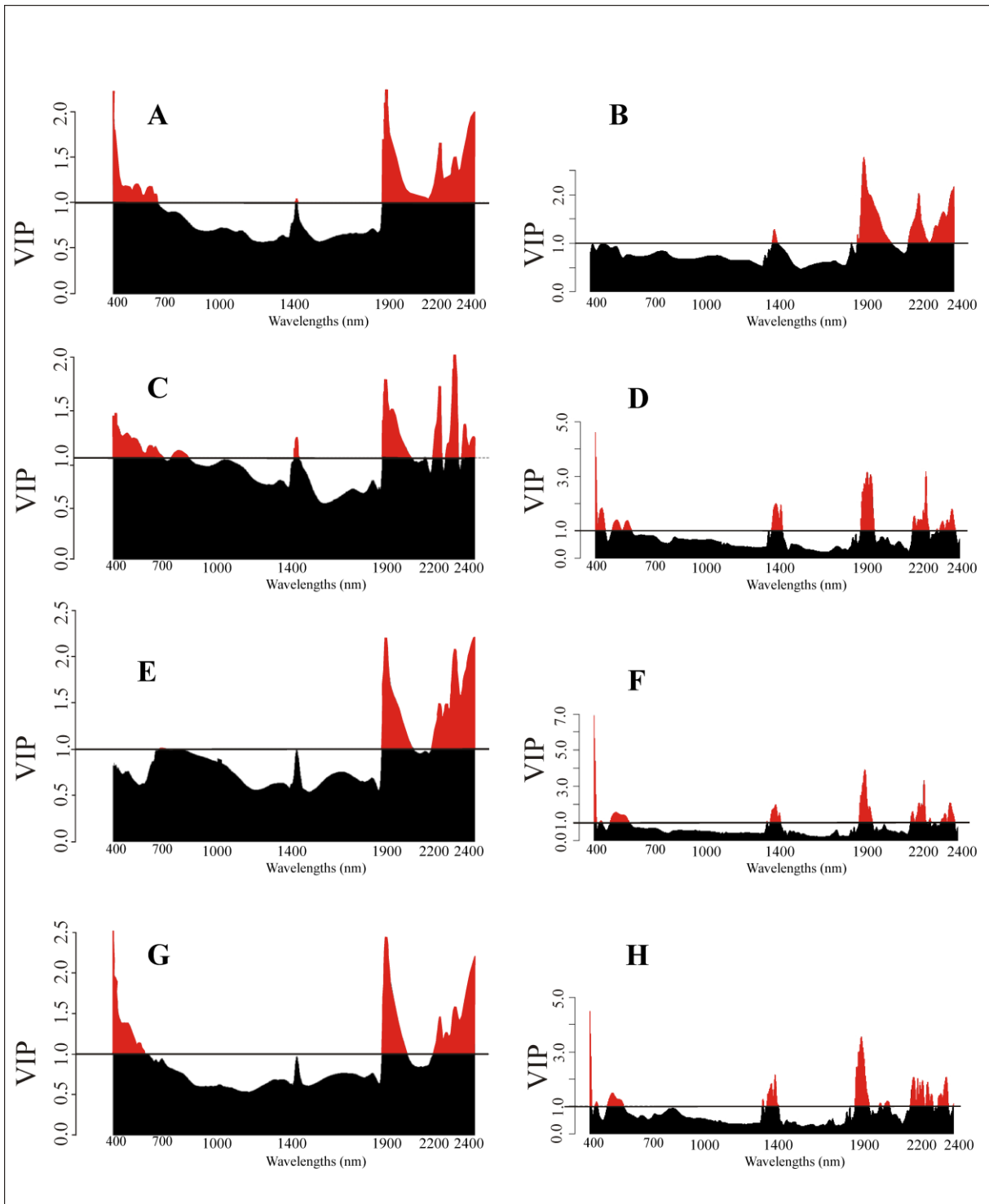
798



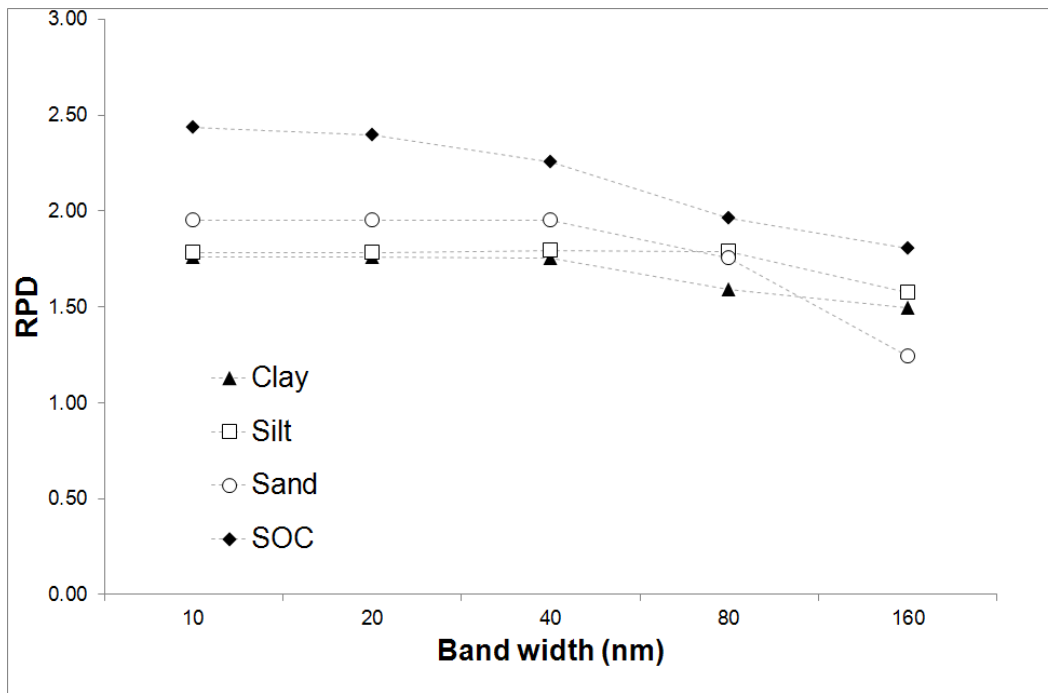
799



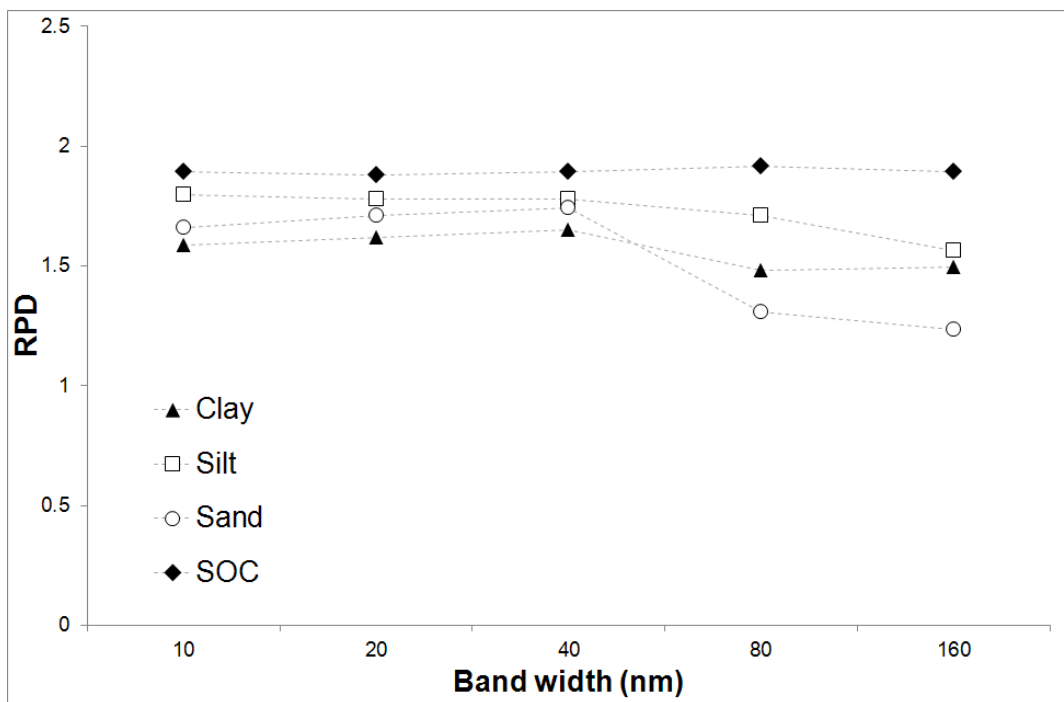
800



801



802



803

804

805 Table 1: Descriptive statistics of soil variables (clay, sand, silt and SOC) of the LUCAS_C
 806 and PONMAC datasets. Columns report basic statistics including the number of samples (n),
 807 standard deviation (sd), coefficient of variation (CV), as well as the type of data
 808 transformation eventually used.

809

810

811

812 Table 2: Main technical characteristics of the multispectral imagers considered in this study.
 813 The FWHM column reports the minimum and the maximum of the imager bandwidths. The
 814 SNR column reports the signal to noise ratio values and the wavelength at which it was
 815 calculated. The SNR condition column reports the radiance values under which the SNR were
 816 calculated.

817

Imager	Spectral bands	Spectral range	FWHM (nm)	SNR	SNR condition
EO-1 ALI	7	4 VNIR 3 SWIR	20-200	572 @550 nm 1040 @1550 nm 912 @2080 nm	17,08 mW/cm ² sr μm 2,15 mW/cm ² sr μm 0,68 mW/cm ² sr μm
LANDSAT 8 OLI	7	5 VNIR 2 SWIR	20-200	100 @562 nm 100 @1610 nm 100 @2200 nm	30 W/m ² sr μm 4,0 W/m ² sr μm 1,7 W/m ² sr μm
Sentinel-2 MSI	12	9 VNIR 3 SWIR	10-60	168 @560 nm 100 @1610 nm 100 @2190 nm	128 W/m ² sr μm 4,0 W/m ² sr μm 1,5 W/m ² sr μm

818

819

820

821

822 Table 3: Main technical characteristics of the hyperspectral imagers considered in this study.
 823 The SNR column reports the signal to noise ratio values and the wavelength at which it was
 824 calculated. The SNR condition column reports the illumination condition used for the SNR
 825 calculation.

826

Imager	Spectral bands	Spectral range (nm)	FWHM (nm)	SNR	SNR condition
Hyperion	220	400-2500	10	161 @550 nm 147 @700 nm 110 @1125 nm 40 @2125 nm	nadir looking 60° sun-zenith angle 0.3 earth albedo
EnMAP	242	420-2450	10	> 500 @495 nm > 150 @2200 nm	nadir looking 30° sun-zenith angle 0.3 earth albedo
PRISMA	247	400-2500	7÷11	600 @ 0.65 μm > 400 @ 1.55 μm > 200 @ 2.1 μm	nadir looking 30° sun-zenith angle 0.3 earth albedo
HypIRI	214	380-2510	10	560 @500 nm 356 @1500nm 236 @2200 nm	nadir looking 23,5° sun-zenith angle 0.25 earth albedo

827

828

829 Table 4: Noisy bands falling into the atmospheric gases absorption ranges that were removed
830 before processing.

Imager	Wavelengths (nm)
Hyperion	356-397; 1134; 1356-1477; 1810-1971; 2405-2577
HypIRI	380-390; 1360-1420; 1820-1950; 2410-2500
EnMAP	1346-1464; 1813-1951; 2408-2438
PRISMA	1352-1428; 1806-1955; 2403-2500
Landsat 8	1370
ALI	none
Sentinel-2	none

831

832

833

834

835

836

837 Table 5: Noise-free results obtained from PONMAC spectral library (*full spectrum* and
838 *resampled* datasets). Mean values of the multiple jack-knifing validation results of soil texture
839 and SOC estimation models. The statistics used are: RMSE = Root Mean Square Error; RPD
840 = Ratio of the Performance to Deviation; RPIQ = Ratio of Performance to Interquartile Range
841 and coefficient of determination (R^2). Different letters within the LSD (least significant

842 difference) column denote a significant difference ($P < 0.05$) among RPIQ values according
 843 to Fisher's LSD's test.

Variable	Resampling type	Spectra transformation ^a	PLSR components	RMSE (%)	RPD	RPIQ	R ²	LSD
Clay	Full spectrum	A+SG	10	4.82	1.97	3.07	0.71	a
	Hyperion	None	11	5.95	1.62	2.42	0.58	c
	HyspIRI	None	12	6.02	1.62	2.43	0.56	bc
	EnMAP	None	11	6.11	1.59	2.39	0.56	c
	PRISMA	None	12	5.91	1.64	2.43	0.58	bc
	Landsat 8	None	5	6.97	1.38	2.07	0.43	e
	ALI	None	7	6.40	1.50	2.25	0.52	d
	Sentinel-2	None	8	5.72	1.70	2.54	0.61	b
Sand	Full spectrum	None	12	5.38	1.83	2.51	0.69	ab
	Hyperion	None	12	5.36	1.86	2.55	0.67	a
	HyspIRI	A	11	5.54	1.79	2.45	0.66	b
	EnMAP	None	12	5.27	1.87	2.56	0.69	a
	PRISMA	None	13	5.43	1.82	2.50	0.67	ab
	Landsat 8	A	6	8.60	1.15	1.56	0.17	e
	ALI	A	8	7.15	1.38	1.88	0.44	d
	Sentinel-2	None	9	5.96	1.66	2.28	0.60	c
Silt	Full spectrum	None	8	4.77	1.95	3.17	0.71	a
	Hyperion	A+SG	8	5.35	1.72	2.77	0.64	c
	HyspIRI	A+SG+D	7	5.11	1.83	2.95	0.66	b
	EnMAP	A	8	5.08	1.82	2.97	0.66	b
	PRISMA	A+SG+D	6	5.08	1.82	2.95	0.67	b
	Landsat 8	A	6	5.87	1.56	2.54	0.55	e
	ALI	A	7	5.60	1.64	2.69	0.60	cd
	Sentinel-2	A	7	5.82	1.58	2.56	0.58	de
SOC	Full spectrum	A+SG	13	0.20	1.78	3.12	0.65	a
	Hyperion	A+SG	19	0.22	1.70	2.95	0.62	bc
	HyspIRI	None	12	0.22	1.65	2.86	0.60	c
	EnMAP	A+SG+D	13	0.20	1.80	3.11	0.67	a
	PRISMA	A	14	0.21	1.75	3.02	0.65	ab
	Landsat 8	None	6	0.25	1.46	2.51	0.50	e
	ALI	None	8	0.25	1.43	2.46	0.49	e
	Sentinel-2	A	7	0.23	1.55	2.68	0.56	d

844 ^a data were transformed using: absorbance (A), first derivative (D) and/or the Savitzky-Golay (SG)
 845 smoothing filter

846

847

848 Table 6: Noise-free results obtained from LUCAS_C spectral library (*full spectrum* and
 849 *resampled* datasets). Mean values of the multiple jack-knifing validation results of soil texture

850 and SOC estimation models. The statistics used are: RMSE = Root Mean Square Error; RPD
851 = Ratio of the Performance to Deviation; RPIQ = Ratio of Performance to Interquartile Range
852 and coefficient of determination (R^2). Different letters within the LSD column denote a
853 significant difference ($P < 0.05$) among RPIQ values according to Fisher's LSD's test.

Variable	Resampling type	Spectra transformation ^a	PLSR components	RMSE (%)	RPD	RPIQ	R^2	LSD
Clay	Full spectrum	A	18	7.63	1.69	2.34	0.64	bc
	Hyperion	A	16	7.98	1.62	2.29	0.61	c
	HyspIRI	A	18	7.71	1.68	2.38	0.64	b
	EnMAP	A+SG+D	18	7.39	1.75	2.48	0.66	a
	PRISMA	A+SG+D	18	7.78	1.66	2.34	0.63	b
	Landsat 8	A	3	10.10	1.27	1.81	0.38	d
	ALI	A	9	9.87	1.30	1.85	0.40	d
	Sentinel-2	A	6	9.92	1.30	1.84	0.40	d
Sand	Full spectrum	A+SG+D	12	12.88	1.25	1.78	0.37	ab
	Hyperion	A	15	13.35	1.22	1.73	0.32	c
	HyspIRI	A	17	12.93	1.25	1.79	0.37	a
	EnMAP	A+SG+D	17	13.28	1.22	1.75	0.30	bc
	PRISMA	A	17	12.98	1.24	1.78	0.35	ab
	Landsat 8	None	7	15.05	1.07	1.53	0.12	e
	ALI	A	9	14.58	1.10	1.59	0.17	d
	Sentinel-2	A	7	14.58	1.11	1.59	0.16	d
Silt	Full spectrum	A+SG+D	14	11.42	1.22	1.74	0.31	cd
	Hyperion	A+SG+D	18	10.93	1.26	1.81	0.38	b
	HyspIRI	A	18	10.66	1.30	1.85	0.40	a
	EnMAP	A	18	10.63	1.31	1.87	0.39	a
	PRISMA	A	18	10.75	1.29	1.84	0.39	ab
	Landsat 8	A	7	11.66	1.19	1.70	0.28	e
	ALI	A	9	11.27	1.23	1.76	0.33	c
	Sentinel-2	A	7	11.61	1.20	1.71	0.28	de
SOC	Full spectrum	A+SG+D	17	0.42	1.35	1.75	0.42	ab
	Hyperion	A	14	0.46	1.23	1.59	0.27	c
	HyspIRI	A	18	0.41	1.37	1.79	0.44	a
	EnMAP	A+SG+D	18	0.41	1.38	1.79	0.45	a
	PRISMA	A+SG+D	18	0.42	1.33	1.73	0.42	b
	Landsat 8	A	5	0.51	1.09	1.40	0.16	e
	ALI	A	5	0.51	1.09	1.41	0.16	e
	Sentinel-2	A	7	0.50	1.13	1.46	0.20	d

854 ^a data were transformed using: absorbance (A), first derivative (D) and/or the Savitzky-Golay (SG)
855 smoothing filter

856 Table 7: Results obtained from PONMAC spectral library with the addition of noise and
857 atmospheric effects (*simulated* datasets). Mean values of the multiple jack-knifing validation
858 results of soil texture and SOC estimation model. The statistics used are: RMSE = Root Mean

859 Square Error; RPD = Ratio of the Performance to Deviation; RPIQ = Ratio of Performance to
 860 Interquartile Range and coefficient of determination (R^2). Different letters within the LSD
 861 column denote a significant difference ($P < 0.05$) among RPIQ values according to Fisher's
 862 LSD's test.

Variable	Resampling type	Spectra transformation ^a	PLSR components	RMSE (%)	RPD	RPIQ	R^2	LSD
Clay	Hyperion	SG	6	7.34	1.31	1.98	0.39	e
	HypIRI	None	9	6.73	1.43	2.16	0.47	b
	EnMAP	A+SG+D	6	6.12	1.57	2.35	0.57	a
	PRISMA	None	8	6.24	1.54	2.33	0.54	a
	Landsat 8	None	5	6.93	1.38	2.07	0.45	cd
	ALI	None	8	6.58	1.45	2.15	0.51	bc
	Sentinel-2	None	5	7.08	1.35	2.03	0.42	de
Sand	Hyperion	SG	9	8.46	1.17	1.59	0.20	d
	HypIRI	None	12	6.45	1.52	2.10	0.53	b
	EnMAP	A+SG	12	6.17	1.60	2.17	0.57	a
	PRISMA	A+SG	12	6.23	1.58	2.17	0.57	a
	Landsat 8	A	6	9.06	1.08	1.48	0.11	e
	ALI	A	8	7.83	1.25	1.73	0.32	c
	Sentinel-2	None	5	8.79	1.12	1.55	0.13	d
Silt	Hyperion	SG	7	6.53	1.40	2.27	0.46	d
	HypIRI	A+SG	8	5.29	1.73	2.80	0.65	a
	EnMAP	A+SG	8	5.23	1.76	2.85	0.65	a
	PRISMA	A+SG	8	5.40	1.70	2.77	0.63	a
	Landsat 8	A	4	6.16	1.47	2.39	0.53	bc
	ALI	A	8	6.06	1.51	2.44	0.53	b
	Sentinel-2	None	4	6.42	1.43	2.34	0.46	cd
SOC	Hyperion	A+SG	7	0.28	1.30	2.25	0.39	c
	HypIRI	A+SG	12	0.25	1.44	2.49	0.49	a
	EnMAP	None	6	0.25	1.45	2.51	0.50	a
	PRISMA	A+SG	7	0.25	1.45	2.48	0.51	a
	Landsat 8	None	6	0.25	1.44	2.48	0.49	a
	ALI	None	7	0.26	1.37	2.38	0.44	b
	Sentinel-2	None	3	0.29	1.26	2.17	0.36	c

863 ^a data were transformed using: absorbance (A), first derivative (D) and/or the Savitzky-Golay (SG)
 864 smoothing filter

865

866

867

868 Table 8: Results obtained from LUCAS_C spectral library with the addition of noise and
869 atmospheric effects (*simulated* datasets). Mean values of the multiple jack-knifing validation
870 results of soil texture and SOC estimation models. The statistics used are: RMSE = Root
871 Mean Square Error; RPD = Ratio of the Performance to Deviation; RPIQ = Ratio of
872 Performance to Interquartile Range and coefficient of determination (R^2). Different letters
873 within the LSD column denote a significant difference ($P < 0.05$) among RPIQ values
874 according to Fisher's LSD's test.

875

Variable	Resampling type	Spectra transformation ^a	PLSR components	RMSE (%)	RPD	RPIQ	R^2	LSD
Clay	Hyperion	SG+D	7	16.98	0.90	1.27	-1.56 _[rc2]	c
	HyspIRI	A	10	9.14	1.41	2.00	0.49	a
	EnMAP	A+SG	11	9.33	1.38	1.95	0.46	a
	PRISMA	A+SG	7	9.11	1.41	2.00	0.50	a
	Landsat 8	A	5	10.11	1.28	1.81	0.37	b
	ALI	A	8	10.09	1.27	1.80	0.37	b
	Sentinel-2	A	6	10.25	1.26	1.78	0.36	b
Sand	Hyperion	SG+D	4	17.92	0.95	1.37	-0.44	e
	HyspIRI	A	11	14.53	1.11	1.59	0.17	a
	EnMAP	A+SG	12	14.57	1.11	1.59	0.17	ab
	PRISMA	A+SG	7	14.50	1.11	1.59	0.18	ab
	Landsat 8	A	5	15.25	1.06	1.51	0.10	d
	ALI	A	8	14.87	1.09	1.55	0.13	bc
	Sentinel-2	A	3	15.18	1.06	1.52	0.10	cd
Silt	Hyperion	A	5	14.71	0.96	1.37	-0.19	c
	HyspIRI	A	8	11.67	1.19	1.69	0.28	ab
	EnMAP	A+SG	9	11.59	1.20	1.71	0.29	ab
	PRISMA	A+SG	7	11.48	1.21	1.72	0.30	a
	Landsat 8	A	7	11.73	1.18	1.68	0.27	b
	ALI	A	8	11.59	1.20	1.71	0.29	ab
	Sentinel-2	A	4	11.90	1.16	1.67	0.25	b
SOC	Hyperion	SG+D	7	0.53	1.05	1.36	0.06	d
	HyspIRI	A	12	0.48	1.15	1.48	0.23	b
	EnMAP	A+SG+D	9	0.48	1.17	1.51	0.25	ab
	PRISMA	A+SG	7	0.48	1.17	1.52	0.26	a
	Landsat 8	A	7	0.51	1.09	1.40	0.14	c
	ALI	A	7	0.51	1.09	1.41	0.15	c
	Sentinel-2	A	6	0.51	1.09	1.41	0.13	c

876 ^a data were transformed using: absorbance (A), first derivative (D) and/or the Savitzky-Golay (SG)
877 smoothing filter

878

879

880

881 Table 9: Number of bands and signal to noise ratio (SNR) of HypIRI, PRISMA and EnMAP
882 imagers in two spectral ranges (400 ÷ 600 nm and 2000 ÷ 2400 nm).

Spectral range	Imager	Bands (n.)	SNR mean	SNR st.dev.
400 -600 nm	HypIRI	21	546.25	71.10
	PRISMA	25	438.34	107.51
	EnMap	35	536.57	50.19
2000 - 2400 nm	HypIRI	41	218.27	52.49
	PRISMA	52	166.37	48.98
	EnMap	48	146.08	41.64

883

884

885